

1 An analysis of early life history in vermilion snapper, *Rhomboplites aurorubens*, based on diver
2 visual surveys of artificial patch-reefs.

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15 Abstract

16 Densities of age-0 and age-1 vermilion snapper, *Rhomboplites aurorubens*, were
17 compared over a nine-year period (2007 to 2015), based on SCUBA visual estimates of densities
18 on small (1.42 m³) artificial reefs (patch-reefs) in the northern Gulf of Mexico. This time period
19 included years both before and after the Deepwater Horizon oil spill in 2010 and provided an
20 evaluation of the effect of the oil spill on this species. Densities of juvenile vermilion snapper on
21 patch-reefs were also compared to catch (number caught/H) of juvenile vermilion snapper from
22 trawl surveys by the Southeast Area Monitoring and Assessment Program (SEAMAP) that has
23 been used as an index of juvenile density in the Gulf of Mexico. High densities of age-0
24 vermilion snapper in 2009, 2014, and 2015 on patch-reefs indicated years of higher potential
25 year-classes of vermilion snapper. Vermilion snapper densities on patch-reefs were not
26 significantly correlated with bottom temperature or sea surface temperature. The density of age-
27 0 vermilion snapper at an offshore location in 2010 was similar to 2007 before the oil spill and
28 for years after the oil spill. Vermilion snapper were less abundant at an inshore location based
29 on their absence from patch-reefs in 2010, 2011, and 2012. We could not determine if these
30 absences were the result of the 2010 oil spill or some other factor. Declines in densities of
31 juvenile vermilion snapper over winter suggested high natural mortality rates ($M = 3.87$) and
32 there were no indications of density dependence for this species. We did not detect a predation
33 or competitive effect by resident fishes as there were no significant differences in juvenile
34 vermilion sapper densities on reefs that were near (15 m) or far (500 m) from larger reef
35 structure. In 2011, vermilion snapper were present on patch-reefs deployed at an east site and
36 absent on patch-reefs deployed at a center site and a west site. Experimental removals of red
37 snapper and other reef fishes in 2013, did not affect juvenile vermilion snapper densities. No

38 significant relation was detected between age-0 and age-1 densities of vermilion snapper. This
39 indicated that age-0 vermilion snapper were not affected by the presence of older conspecifics, or
40 that age-1 vermilion snapper densities in the present study were too low to influence age-0
41 densities. There was a significant positive correlation between the densities of juvenile (age-0
42 and age-1) vermilion snapper and red snapper, *Lutjanus campechanus*, in September, and
43 between age-0 vermilion snapper density and the densities of age-0 tomtate, *Haemulon*
44 *aurolineatum*, and other reef fishes in October. In 2015, there were higher densities of age-0
45 vermilion snapper on patch-reefs located farther to the east, but it was not determined if this
46 difference was caused by lower initial densities of other resident reef fishes at the east site, or by
47 the location of the patch-reefs. There was a significant correlation between the density of age-0
48 vermilion snapper on patch-reefs in October and catch-per-unit-effort (CPUE = catch/H) of
49 vermilion snapper from SEAMAP fall trawl surveys east of the Mississippi River. However, in
50 June there was no significant correlation between the density of age-1 vermilion snapper on
51 patch-reefs and CPUE from SEAMAP summer trawl surveys. The visual survey methods of
52 patch-reefs yielded significantly different estimates of densities and size of juvenile vermilion
53 snapper when compared to samples taken with drop-net and rotenone. However, the error in size
54 estimation was small and individuals were still assigned to the correct age class from visual
55 surveys. The patch-reef surveys used in this study have potential for use in developing indexes
56 of juvenile vermilion snapper abundance. However, further validations are needed of diver
57 visual estimates of juvenile vermilion snapper on patch-reefs by comparisons with other methods
58 before this method is widely applied.

59 Introduction

60

61 Vermilion snapper support both commercial and recreational fisheries in the Gulf of
62 Mexico. Presently, vermilion snapper are not considered overfished and are not undergoing
63 overfishing (SEDAR-67 2020). However, vermilion snapper stocks may experience higher
64 fishing mortality as restrictions on other reef fish species cause fishers to target vermilion
65 snapper as an alternative (Schirripa 2000; Moncrief et al. 2018). Vermilion snapper are also
66 preyed upon by reef fish such as red snapper and greater amberjack, *Seriola dumerili* (Manooch
67 and Haimovici 1983; Szedlmayer and Brewton 2020). It is possible that vermilion snapper
68 stocks could decline as stocks of their predators recover. Therefore, it is important to monitor
69 vermilion snapper stocks to ensure they remain healthy as these potential sources of mortality
70 increase.

71 Accurate stock assessment and management of marine reef fish benefits from an
72 understanding of juvenile settlement. Management efforts are more effective if year-class
73 strength can be estimated before juveniles enter the fishery, rather than back-calculating year-
74 class strength after a year-class moves into the exploited portion of the fishery. The open nature
75 and large size of marine nursery habitats make accurate measurement of juvenile fish density
76 difficult. Accurately predicting year-class strength could allow quotas to be increased when it
77 can be anticipated that larger year-classes will enter the fishery, and stocks could be protected
78 from overfishing by decreasing quotas as less abundant year-classes enter the fishery. Presently,
79 density estimates of juvenile vermilion snapper in stock assessments are based on trawl surveys
80 (SEDAR-67 2020). However, if juvenile vermilion snapper reside on structured habitat, trawl
81 sampling may be an ineffective sampling technique for this species. Therefore, other survey

82 methods may be more appropriate for determining the density of juvenile vermilion snapper after
83 they settle to benthic habitats and move to reef structure.

84 Small isolated reefs (patch-reefs) have long been used to experimentally manipulate reef
85 fish communities (Sale 1980; Doherty 1982; Steele 1998). Patch-reefs can be easily manipulated
86 and can facilitate experimental designs that address specific ecological questions. Recently,
87 several studies have used artificial patch-reefs to evaluate different aspects of juvenile red
88 snapper and gray triggerfish biology (Simmons and Szedlmayer 2011; Mudrak and Szedlmayer
89 2012, 2020a; Szedlmayer and Mudrak 2014). These studies deployed patch-reefs with identical
90 designs, in the same general area, and at similar time periods each year from 2007 to 2015.

91 The objective of the present study was to re-examine these patch-reef visual surveys to
92 evaluate interannual differences in juvenile vermilion snapper density. This re-examination will
93 produce a measure of density each year that can be used as an index of abundance, for inclusion
94 in vermilion snapper stock assessment efforts. Based on these visual surveys we can estimate
95 natural mortality rates, which are important for modeling vermilion snapper population
96 dynamics. Annual densities will be compared with catch-per-unit-effort (CPUE) from SEAMAP
97 trawl surveys presently used as an index of juvenile vermilion snapper abundance. These annual
98 densities from patch-reefs will also be compared to both sea surface and bottom temperature.
99 We will also compare juvenile vermilion snapper densities to the densities of older conspecifics
100 and other reef fishes. In addition, because data are available for years both before and after the
101 2010 Deepwater Horizon oil spill, we will assess possible effects of this oil spill on juvenile
102 vermilion snapper densities. Finally, we will examine the effects of patch-reef location, distance
103 from larger reef structure, and the densities of red snapper and other reef fishes on vermilion

104 snapper densities that could potentially yield further insights into the early life history of
105 vermilion snapper.

106

107 Methods

108

109 Reef design and surveys

110

111 Each patch-reef had a total volume of 1.42 m³ and consisted of a plastic pallet (1.22 ×
112 1.02 × 0.14 m), 10 concrete blocks (41 × 20 × 10 cm), and a plastic crate (65 × 35 × 28 cm;
113 Figure 1). Patch-reefs were assembled with 122 cm plastic cable ties with a breaking strength of
114 79 kg. Small plastic floats (5.1 × 12.7 cm) were attached to each of the four corners, and a larger
115 float (15.2-cm diameter) was attached to the center of each patch-reef. All floats were attached 1
116 m above the patch-reef. The floats added vertical structure to the patch-reef and facilitated
117 patch-reef relocations with sonar. The patch-reefs were anchored to the seafloor by attachment
118 to a 1.2 m ground anchor with a 3 m length of 1.3 cm diameter nylon rope. All patch-reefs were
119 placed at least 500 m apart and 500 m away from any known reefs in the area (Mudrak and
120 Szedlmayer 2012).

121 Patch-reefs were visually surveyed by SCUBA divers. Divers identified fish to species,
122 counted all fish present and estimated sizes in 25-mm total length (TL) intervals. Divers held a
123 stationary position 2 m from the patch-reef and counted all fish within visible range of the patch-
124 reef over an approximate 15-min survey period. Fish distances varied and were not measured,
125 and all densities were calculated as density per m³ reef size. However, diver visibility typically
126 exceeded the maximum fish distances from the reef due to the small size of the patch-reefs. If

127 diver visibility was determined to be less than the 3 m distance to the far side of the reef (i.e.,
128 divers could not count all fish on the far side of the reef) the reef surveys were discontinued.
129 Some of the patch-reefs became partially buried after storms. If more than 50 % of a patch-reef
130 was buried, the estimate of fish density from that patch-reef was not included in the analysis.
131 The age of vermilion snapper observed was estimated based on TL as determined by a two factor
132 Von Bertalanffy growth equation reported by Moncrief et al. (2018). All vermilion snapper
133 greater than 279 mm TL were considered age-2 or older. Vermilion snapper were considered
134 age-0 in May, June, and July when less than 76 mm TL, in August when less than 102 mm TL,
135 in September when less than 127 mm TL, and in October, November, and December when less
136 than 152 mm TL. No surveys were conducted in January, February, March, or April. At the
137 time of the diver surveys, temperature, salinity, and dissolved oxygen were measured within 1 m
138 of the seafloor with a remote YSI 6920 meter. If more than one environmental measurement was
139 taken at a reef site during a survey, we used the mean values temperature, salinity, and dissolved
140 oxygen for analyses. Temperature ranged from 22.9 to 30.0 °C, salinity from 29.0 to 36.2 ppt
141 and dissolved oxygen (DO) from 2.0 to 6.5 ppm (Table 1).

142

143 Annual comparisons

144

145 The densities (number of fish/m³ of patch-reef size) of age-0 and age-1 vermilion snapper
146 were compared among deployment dates, locations, and survey dates (Table 2). Patch-reefs
147 deployed at the same time and location were referred to as a reef set (Table 2; Figure 2). The
148 patch-reefs (described above) were deployed with 10 to 30 patch-reefs per set. One set of patch-
149 reefs was deployed each year, with the exception of 2010 when three patch-reef sets ($N = 10$

150 patch-reefs for each set, $N = 30$ total patch-reefs) were deployed to evaluate the effect of the
151 Deepwater Horizon oil spill on reef-associated fish assemblages (Table 2). The offshore location
152 was 19 – 23 km from shore and ranged in depth from 17 – 24 m (Figure 2). The inshore location
153 was 12 – 16 km from shore and ranged in depth from 14 – 18 m (Figure 2). If there was more
154 than one survey in the same month, the highest mean density of age-0 vermilion snapper per
155 survey was used for annual comparisons of juvenile fish densities. In 2008, all patch-reefs were
156 lost after Hurricane Gustav (1 September 2008). In 2009, patch-reefs were lost or damaged after
157 Hurricane Ida (10 November 2009). In 2011, one patch-reef was lost after tropical storm Lee (4
158 September 2011), and in 2012 four patch-reefs were lost after Hurricane Isaac (28 August 2012).

159 Patch-reefs were deployed with experimental designs to examine the effects of proximity
160 to larger reefs, spatial distribution, and the addition or removal of potential predators and
161 competitors (Simmons and Szedlmayer 2011; Mudrak and Szedlmayer 2012; Szedlmayer and
162 Mudrak 2014). However, for comparing densities among years we only used fish densities on
163 patch-reefs that were deployed in July or August, placed at least 500 m from other known reefs,
164 and without fish artificially added or removed from a patch-reef.

165 Diver visual data allowed comparisons of vermilion snapper densities among years in
166 four months (Table 2). The density of vermilion snapper observed in August included data from
167 eight years (2008 to 2015), in September from five years (2007, 2009, 2010, 2012, and 2014), in
168 October from six years (2007, 2010, 2011, 2013, 2014, and 2015) and in June from six years for
169 patch-reefs that were deployed the previous year (2007, 2010, 2011, 2013, 2014, and 2015).

170 The Deepwater Horizon oil spill occurred from 20 April to 15 July 2010 (NOAA 2010;
171 Allan et al. 2012), and was predicted to affect local fish populations (Rooker et al. 2013). In
172 2010, 10 patch-reefs were deployed in July at an offshore location (Off-Jul2010) that was the

173 same location as the 2008 and 2009 patch-reefs (Mudrak and Szedlmayer 2012; Figure 2). Two
174 additional patch-reef sets (each with $N = 10$) were deployed in August 2010. The Off-Aug2010
175 reef set was deployed at the same offshore location as the Off-Jul2010 reef set, and the In-
176 Aug2010 reef set was placed closer to shore (Figure 2). Vermilion snapper densities from these
177 three reef sets in 2010 were analyzed separately when comparing the effect of interannual
178 differences in density, because differences in location and deployment date could be associated
179 with differences in the density of vermilion snapper (Szedlmayer and Mudrak 2014). All reef
180 sets after 2010 were deployed at the inshore study location (Figure 2).

181

182 Age-0 and age-1 vermilion snapper interactions

183

184 We compared the densities of age-0 and age-1 vermilion snapper on patch-reefs with
185 Pearson correlation analysis, but only used densities from patch-reefs that were not manipulated
186 (i.e. deployed in July or August at the offshore or inshore locations at least 500 m from other
187 reefs, with no fish added or removed). Surveys from August, September, and October were
188 analyzed separately, and June surveys were not compared because no age-0 vermilion snapper
189 were observed.

190

191 Interactions with other species

192

193 Densities of vermilion snapper were compared to the density of other reef fish species
194 residing on unmanipulated patch-reef. For August, September, and October, we used partial
195 correlations to compare the density of juvenile (age-0 and age-1) vermilion snapper with the

196 density juvenile red snapper (age-0 and age-1) with the effects of other reef fish removed. In
197 June, partial correlations were used to compare the density of all vermilion snapper with the
198 density all red snapper with the effects of other reef fish removed. Other reef fish was defined
199 here as all fish counted in visual surveys except for red snapper and open habitat or pelagic
200 species. Open habitat or pelagic species that were observed but not included in patch-reef
201 density estimates included Atlantic bumper, *Chloroscombrus chrysurus*, blue runner, *Caranx*
202 *crysos*, flounder, *Paralichthys sp.*, grass pogy, *Calamus arctifrons*, king mackerel,
203 *Scomberomorus cavella*, lizardfish, Synodontidae, longspine pogy, *Stenotomus caprinus*,
204 lookdown, *Selene vomer*, round scad, *Decapterus punctatus*, searobin, *Prionotus sp.*, Spanish
205 mackerel, *Scomberomorus maculatus*, and spot, *Leiostomus xanthurus*.

206 In August, September, and October, we used partial correlations to compare the density
207 of juvenile vermilion snapper to the density of juvenile tomtate, *Haemulon aurolineatum*, with
208 the effect of other reef fish removed. Other reef fish was defined here as all fish counted in
209 visual surveys except for tomtate and open habitat or pelagic species. In June, the density of all
210 age classes of vermilion snapper and tomtate were compared with the effect of all other reef fish
211 removed. The June survey was before recruitment of age-0 individuals of either species to the
212 patch-reefs. Age-0 tomtate were defined as individuals less than 102 mm in August, September,
213 and October based on a Von Bertalanffy growth curve relation (Norberg 2015).

214 In August, September, October, and June we used correlation analysis to compare the
215 density of juvenile vermilion snapper to the total density of all other reef fish. Other reef fish
216 was defined here as all fish counted in visual surveys except red snapper, vermilion snapper and
217 open habitat or pelagic species.

218

219 Environmental correlations

220

221 Bottom temperature was measured with temperature loggers (U22-001, Onset
222 Incorporated) deployed at one of three stations over the time period of the present study. These
223 stations were located 31 – 32 km southeast of Dauphin Island Alabama U.S., at depths of 26 – 30
224 m. Sea surface temperatures were obtained from the 42012 data buoy located 81 km southeast of
225 Mobile Alabama U.S. (NOAA 2020). For comparisons with age-0 vermilion snapper densities
226 each month, and age-1 vermilion snapper densities in June, we used the mean of all bottom or
227 sea surface temperatures for each month.

228

229 Mortality estimates

230

231 We used density estimates of vermilion snapper from patch-reefs deployed in July or
232 August that were at least 500 m from larger reef structures and without fish experimentally
233 added or removed for mortality estimates. In addition to the offshore and inshore locations, this
234 included reefs located at the West, East and 2015 East sites (Figure 2). To be used in mortality
235 estimates, the patch-reef needed a fall visual survey and a visual survey the next summer. A
236 total of 127 patch-reefs fit these criteria. The highest density of age-0 vermilion snapper
237 observed on individual patch-reefs in the fall were used to calculate a mean density of age-0
238 vermilion snapper each year. The first survey of these same patch-reefs the next summer was
239 used to calculate a mean density of age-1 vermilion snapper. Survival (S) was calculated as the
240 mean density of age-1 vermilion snapper on patch-reefs in the summer divided by the mean
241 density of age-0 vermilion snapper observed the previous fall for each year. Total instantaneous

242 mortality (Z), was calculated as $Z = -\ln(S^{365/t})$, where t = the mean number of days between the
243 fall and summer surveys for each particular year. Mean Z was calculated as the mean of the
244 annual mortalities observed each year. However, in 2011 and 2013 mean densities of vermilion
245 snapper increased over winter causing annual survival rates greater than one. Therefore, Z
246 estimates from 2011 and 2013 were undefined and not included in further analyses.

247

248 Comparison to SEAMAP trawl surveys

249

250 The densities of vermilion snapper from diver visual surveys of patch-reefs were
251 compared to the catch-per-unit-effort (CPUE = catch/H) of vermilion snapper estimated from
252 trawl surveys (Southeast Area Monitoring and Assessment Program – SEAMAP; Gulf States
253 Marine Fisheries Commission 2018). For comparisons, we only used SEAMAP trawl surveys
254 that were taken during the same time periods as the patch-reef visual surveys. These included
255 SEAMAP trawl surveys in June and October of each year. Also, we only used SEAMAP trawl
256 surveys that were taken east of the Mississippi River ($> -89^\circ$ W). Most vermilion snapper
257 collected by trawl were measured by fork length (mm FL), and these were converted to TL with
258 the equation $TL = 1.128 \times FL - 2.112$ (Moncrief et al. 2018). We applied the same TL to age
259 relation that was used to estimate age for fish from patch-reefs, to estimate age from length for
260 vermilion snapper collected by trawl. The mean CPUE of age-0 vermilion snapper from October
261 trawls was compared to the mean density estimates of age-0 vermilion snapper on patch-reefs in
262 October. The mean CPUE of age-1 vermilion snapper from June trawls was compared to the
263 mean density estimates of age-1 vermilion snapper on patch-reefs in June. For the comparison of

264 trawl CPUE to patch-reef surveys in 2010, we used a mean density of juvenile vermilion snapper
265 derived from both the Off-Aug2010 and In-Aug2010 reef sets.

266

267 Effects of distance from larger reef structure

268

269 In 2008, 2009, and 2010, we examined the effect of patch-reef proximity to larger
270 artificial reefs on the density of juvenile fish on patch-reefs (Mudrak and Szedlmayer 2012).
271 Each year we deployed 10 patch-reefs 15 m (Near) from larger steel cage artificial reefs (2.5 x
272 1.3 x 2.4 m) and 10 patch-reefs at a distance of 500 m (Far) from large steel cage artificial reefs
273 for the Off-Jul2008, Off-Jul2009, and Off-Jul2010 deployments. Both the Near and Far patch-
274 reefs were deployed and surveyed at the same time each year, which allowed for comparisons of
275 juvenile vermilion snapper densities on patch-reefs in areas used by predators and competitors to
276 densities on patch-reefs away from known sources of predators and competitors. One Near
277 patch-reef in August 2009 had densities of age-0 vermilion snapper too large for divers to count.
278 For this patch-reef, photographs were used to estimate densities of age-0 vermilion snapper. All
279 age-0 fish were counted in a subsampled portion of the photo and extrapolated to estimate the
280 total number of fish in the photo. A second close-up photo of the patch-reef was used to estimate
281 the proportion of vermilion snapper among the identifiable fish.

282

283 2011 spatial experiment

284

285 In 2011, there were three patch-reef sets deployed at a Center site, a West site and an East
286 site. The Center site (In-Jul2011, $N = 10$) was 13 km south of the coastline (30.107°N,

287 87.958°W), the West site ($N = 10$) was 30 km west of the center site and the East site ($N = 10$)
288 was 30 km east of the center site (Figure 2). All three 2011 patch-reef sets were deployed and
289 surveyed at similar times and allowed for comparisons of juvenile vermilion snapper densities at
290 larger spatial scales. Over the winter, eight patch-reefs were lost at the West site, possibly due to
291 shrimp trawling. Therefore, the density estimates of vermilion snapper on the remaining two
292 patch-reefs at the West site were not used for analysis in June due to low sample size. This
293 deployment of East, Center, and West patch-reefs was repeated in 2012, but Hurricane Isaac
294 buried many of the patch-reefs at the East and West sites on 28 August 2012, and sample sizes
295 were too low for spatial comparisons in 2012.

296

297 2015 spatial experiment

298

299 In 2015, it was possible that the 100 patch-reefs that were previously deployed at the
300 inshore location over the years since 2010 were providing a source of age-1 red snapper, gray
301 triggerfish, and other reef fish that could quickly colonize any new reefs built in the immediate
302 area. These age-1 individuals may affect the density of age-0 individuals (Mudrak and
303 Szedlmayer 2012; Szedlmayer and Mudrak 2014). To examine this possibility two patch-reef
304 sets were deployed in 2015. One patch-reef set was deployed at the center site (In-Jul2015, $N =$
305 15) 500 m from previous patch-reef deployments, and one patch-reef set (2015 East site, $N = 15$)
306 was deployed 11 km east of the center site (Figure 2). These two patch-reef sets were deployed
307 and surveyed at similar times in 2015. The 2015 East site was selected so that all patch-reefs
308 were at least 1 km from other known reefs in the area. Placing patch-reefs 11 km to the east of

309 the center site patch-reefs allowed for comparisons of vermilion snapper densities both with
310 (center) and without (east) a near-by source of immigrants.

311

312 Removal experiment

313

314 In 2013, we applied a removal experiment to examine the effects of age-1 red snapper
315 and other reef associated fish species on age-0 vermilion snapper densities. In June, 30 patch-
316 reefs were deployed at the center site (June patch-reefs). Fish were able to colonize these June
317 patch-reefs for one month prior to the start of the removal treatments. In July, 10 of these June
318 patch-reefs had all fish removed (All Removed). Scuba divers placed a 3-m radius cast net
319 (drop-net) over the reef and buried the lead line in the sand. Scuba divers then dispensed
320 rotenone onto the patch-reef and collected all fish in the net. For a red snapper only removal (RS
321 Removal), 10 of the June patch-reefs had only red snapper removed with fish traps on 6 and 8
322 August 2013. The traps ($1.2 \times 1.5 \times 0.6$ m; Collins 1990) were baited with squid *Loligo* sp., and
323 gulf menhaden *Brevoortia patronus*. The trap was set next to (< 5 m) each patch-reef for 15
324 minutes before retrieval. All captured red snapper were removed from the patch-reef, while all
325 other captured fish were immediately released at the patch-reef site. The other 10 patch-reefs
326 deployed in June 2015 served as a control with no removals (Control Reefs).

327 In July, we also deployed 10 new patch-reefs and defined these as a “New Reef”
328 treatment (In-Jul2013). These New Reefs served as empty patch-reefs with few if any resident
329 fish, because there was little time for fish recruitment before they were surveyed. These
330 manipulations of patch-reefs with removals and non-removal allowed for determinations of the
331 effects of resident reef fish on the recruitment of age-0 vermilion snapper.

332

333 Comparisons of visual surveys to drop-net-rotenone sampling

334

335 Patch-reefs ($N = 14$) were first visually surveyed prior to drop-net-rotenone collections.
336 Immediately after visual surveys, drop-net-rotenone collections were carried out. These visual
337 and drop-net-rotenone surveys were completed on four patch-reefs in November 2012 and 10
338 patch-reefs in July 2013. All fish collected with drop-nets were placed on ice and returned to the
339 laboratory for analysis. In the laboratory all fish were identified to species, weighed (nearest 0.1
340 g) and lengths measured (standard length, FL, TL mm). These visual surveys followed by drop-
341 net-rotenone collections allowed for validation of visual surveys methods used to estimate the
342 density and size of vermilion snapper on patch-reefs.

343

344 Statistical analysis

345

346 Annual densities of juvenile vermilion snapper were examined for possible effects of the
347 various treatments with generalized linear models (GLIMMIX; SAS 9.4) with negative binomial
348 distributions and logarithm-link functions (Huelsenbeck and Crandall 1997; Seavy et al. 2005;
349 Bolker et al. 2009). Due to the many zero density counts on patch-reefs, one was added to the
350 vermilion snapper density on all patch-reefs being compared. If significant differences were
351 detected among densities, specific differences were identified with a Tukey multiple comparison
352 test (Zar 2010). In our statistical analyses, annual comparisons of mean densities were analyzed
353 with separate tests for August, September, October, and June (i.e., not all reef sets were surveyed
354 all months analyzed). However, when comparing the effects of differences in distance from

355 larger reef structure, patch-reef locations, and fish removals we used a repeated measures design
356 (RM).

357 A Pearson's product-moment correlation coefficient was calculated to determine the
358 association between the CPUE from trawls to densities on patch-reefs from visual surveys, to
359 compare densities of age-0 and age-1 vermilion snapper, and to compare vermilion snapper
360 densities with temperature. In addition, to determine if density dependent mechanisms were
361 occurring, we compared mean age-0 vermilion snapper densities in the fall to total mortality (Z)
362 each year. Partial correlation was used to compare densities of vermilion snapper, red snapper,
363 and other reef fishes. Drop-net-rotenone samples were compared to visual estimates with a
364 Fisher's exact test. All statistical differences were considered significant at $P \leq 0.05$

365

366 Results

367

368 Annual variation of juvenile vermilion snapper density on patch-reefs

369

370 The density of age-0 vermilion snapper observed on small artificial patch-reefs in the fall
371 varied among years, and few age-1 vermilion snapper remained on patch-reefs the following
372 summer (Figure 3). The density of age-0 vermilion snapper was significantly different among
373 years in August ($F_{9,99} = 8.29, P < 0.001$; Figure 4), September ($F_{6,63} = 13.7, P < 0.001$; Figure 5),
374 and October ($F_{6,70} = 17.2, P < 0.001$; Figure 6). The density of age-1 vermilion snapper was not
375 significantly different among years in August ($F_{79,99} = 0.22, P = 0.991$; Figure 4) or September
376 ($F_{6,63} = 1.26, P = 0.287$; Figure 5). There were significant differences in age-1 vermilion

377 snapper densities among years in October ($F_{6,70} = 2.94$, $P = 0.013$; Figure 6), and in June ($F_{6,84} =$
378 5.92 , $P < 0.001$; Figure 7).

379

380 Age-0 and age-1 vermilion snapper

381

382 There were no significant correlations between age-0 and age-1 vermilion snapper
383 densities in August ($r = -0.01$, $P = 0.935$, $N = 89$ patch-reefs), September ($r = -0.04$, $P = 0.749$, N
384 $= 70$ patch-reefs) or October ($r = -0.03$, $P = 0.766$, $N = 77$ patch-reefs).

385

386 Correlations with other species

387

388 A total of 57 species of reef fish were counted in the present study. Species that
389 comprised more than 1% of the total fish counted include red snapper (34.4%), tomtate (17.9%),
390 pigfish, *Orthopristis chrysoptera*, (12.1%), gray triggerfish (11.5%), vermilion snapper (6.3%),
391 rock sea bass, *Centropristis philadelphica*, (4.5%), Atlantic spadefish, *Chaetodipterus faber*,
392 (3.1%), sand perch, *Diplectrum formosum*, (2.5%), lane snapper, *Lutjanus synagris*, (2.4%), and
393 pygmy filefish, *Stephanolepis setifer*, (1.6%).

394 Juvenile (age-0 and age-1) vermilion snapper density was not significantly correlated
395 with juvenile red snapper density in August ($r = -0.087$, $P = 0.369$, $N = 109$ patch-reefs), or
396 October ($r = -0.08$, $P = 0.517$, $N = 77$ patch-reefs). There was a significant correlation between
397 juvenile vermilion snapper densities and juvenile red snapper densities in September ($r = 0.28$, P
398 $= 0.020$, $N = 70$ patch-reefs). The density of all vermilion snapper was not significantly
399 correlated with the density of all red snapper in June ($r = 0.04$, $P = 0.703$, $N = 91$ patch-reefs).

400 Juvenile vermilion snapper density was not significantly correlated with juvenile tomtate
401 density in August ($r = -0.01$, $P = 0.907$, $N = 109$ patch-reefs), or September ($r = 0.02$, $P = 0.862$,
402 $N = 70$ patch-reefs). There was a significant correlation between juvenile vermilion snapper and
403 juvenile tomtate densities in October ($r = 0.38$, $P < 0.001$, $N = 77$ patch-reefs). The density of all
404 vermilion snapper was not significantly correlated with the density of all tomtate in June ($r = -$
405 0.02 , $P = 0.832$, $N = 91$ patch-reefs).

406 The density of juvenile vermilion snapper was not significantly correlated with total reef
407 fish density in August ($r = -0.08$, $P = 0.434$, $N = 109$ patch-reefs), or September ($r = -0.06$, $P =$
408 0.649 , $N = 70$ patch-reefs). There was a significant correlation between juvenile vermilion
409 snapper density and total reef fish density in October ($r = 0.42$, $P < 0.001$, $N = 77$ patch-reefs).
410 The density of all vermilion snapper density was not significantly correlated with total reef fish
411 density in June ($r = 0.016$, $P = 0.879$, $N = 91$ patch-reefs).

412

413 Environmental conditions

414

415 Bottom temperature data were available for most of the present study except November
416 2008 – September 2010. There were no significant correlations between bottom temperature and
417 age-0 vermilion snapper density in August ($r = 0.25$, $P = 0.631$, $N = 6$), September ($r = -0.40$, P
418 $= 0.735$, $N = 3$), or October ($r = 0.05$, $P = 0.922$, $N = 6$). There was no significant correlation
419 between bottom temperature and age-1 vermilion snapper density in June ($r = 0.50$, $P = 0.314$, N
420 $= 6$).

421 Sea surface temperature data were available for all years beginning in 2009. There was
422 no significant correlation between monthly mean sea surface temperature and age-0 vermilion

423 snapper density in August ($r = -0.47$, $P = 0.291$, $N = 7$), September ($r = -0.25$, $P = 0.754$, $N = 4$),
424 or October ($r = -0.10$, $P = 0.867$, $N = 5$), or between monthly mean sea surface temperature and
425 mean age-1 vermilion snapper density in June ($r = -0.75$, $P = 0.146$, $N = 5$).

426

427 Mortality

428

429 Among the patch-reefs deployed in July or August that did not have fish experimentally
430 added or removed, there were 127 patch-reefs that were surveyed both in the fall and the
431 following summer. The time between the fall survey with the highest mean age-0 vermilion
432 snapper density and the first survey the next summer ranged from 189 to 353 days with a mean
433 of 263 days between surveys. Vermilion snapper densities increased over winter in 2011 and
434 2013, and mortality rates were not estimated. Mortality (Z) in years with observed declines in
435 density ranged from a low of $Z = 2.69$ in 2014 to a high of $Z = 4.76$ in 2010 (Table 3). There
436 was no significant relation between mean age-0 vermilion snapper density each fall and Z for
437 that year ($r = 0.464$, $P = 0.3537$, $N = 6$).

438

439 Comparison to SEMAP trawl surveys

440

441 There was a significant positive correlation between mean age-0 vermilion snapper
442 densities on patch-reefs and mean CPUE in SEMAP trawl surveys each October ($r = 0.93$, $P =$
443 0.008 ; Table 4; Figure 8). There was no significant correlation between age-1 vermilion snapper
444 densities on patch-reefs and CPUE in SEMAP trawl surveys in June ($r = -0.17$, $P = 0.748$; Table
445 4; Figure 9).

446

447 Distance from Larger Reef Structure

448

449 There was no significant difference in age-0 vermilion snapper densities on patch-reefs
450 that were Near or Far from larger reef structure in August (mean \pm SE; Near = 39.72 ± 38.90 , $N = 30$;
451 Far = 3.62 ± 3.52 , $N = 30$) or September (Near = 51.37 ± 35.69 , $N = 20$; Far = $51.37 \pm$
452 35.24 , $N = 20$), and this pattern persisted with age-1 densities the following July (Near = 0, $N = 5$;
453 Far = 0.14 ± 0.14 , $N = 5$; RM $F_{1,58} = 0.44$, $P < 0.512$). There was no significant difference in
454 age-1 vermilion snapper densities on patch-reefs that were Near or Far from larger reef structure
455 in August (Near = 0.02 ± 0.02 , $N = 30$; Far = 0.16 ± 0.12 , $N = 30$), or in September (Near = 0.14
456 ± 0.08 , $N = 20$; Far = 0.11 ± 0.08 , $N = 20$, RM $F_{1,58} = 1.04$, $P = 0.312$).

457

458 Spatial distribution of reefs

459

460 Densities of the 2011 year class (age-0 in 2011, age-1 in 2012) of vermilion snapper on
461 patch-reefs deployed in 2011 were significantly different among locations, with the highest
462 densities at the East site, and no vermilion snapper present at the Center or West sites (RM $F_{2,27}$
463 = 50.9, $P < 0.001$; Figure 10). Densities of the 2010 year class (age-1 in 2011) of vermilion
464 snapper on patch-reefs deployed in 2011 were not significantly different among locations
465 (RM $F_{2,27} = 3.1$, $P = 0.061$; Figure 11).

466

467 2015 spatial distribution

468

469 Densities of the 2015 year class (age-0 in 2015, age-1 in 2016) of vermilion snapper on
470 patch-reefs deployed in 2015 were significantly different among locations ($F_{1,28} = 62.6, P <$
471 0.001), time ($F_{2,55} = 10.7, P < 0.001$), and location x time interaction ($F_{2,55} = 6.28, P = 0.004$;
472 Figure 12). Age-0 vermilion snapper densities were significantly higher on the 2015 East site
473 compared to the center site and remained higher the following year in June (Figure 12).
474 Densities of the 2014 year class (age-1 in 2015) of vermilion snapper on patch-reefs deployed in
475 2015 were significantly affected by location ($F_{1,28} = 53.8, P < 0.001$), and time ($F_{1,28} = 9.26, P =$
476 0.005), but not by a location x time interaction ($F_{1,28} = 3.7, P = 0.066$, Figure 13).

477

478 Removal experiment

479

480 Densities of the 2013 year class (age-0 in 2013, age-1 in 2014) of vermilion snapper on
481 patch-reefs deployed in 2013 were not significantly affected by removal treatments (RM $F_{3,36} =$
482 $0.27, P = 0.843$; Figure 14). There was also no significant difference in densities of the 2012
483 year-class (age-1 in 2013) of vermilion snapper on the 2013 patch-reefs (RM $F_{3,36} = 0.14, P =$
484 0.938 ; Figure 15).

485

486 Drop-net-rotenone sampling

487

488 Among the 14 patch-reefs that were surveyed visually and with drop-net-rotenone
489 collections, 10 had at least one vermilion snapper detected by one or both survey methods, and
490 all patch-reefs had less than 35 individuals detected by either method. Among the 10 patch-reefs
491 with detected vermilion snapper, three had the same number of individuals for both methods

492 (these three patch-reefs only had one individual present), four had counts that differed by one
493 individual between the two methods, two differed by two individuals, and one differed by 10
494 individuals. A total of 59 vermilion snapper were counted in the visual survey, and 63 vermilion
495 snapper were captured by drop-nets. Comparison of the 52 individuals detected by both methods
496 indicated that 46 % ($N = 24$) had measured TL that matched the 25 mm size interval visually
497 estimated by divers, 0 % had estimated size interval smaller than their measured length, and 54
498 % ($N = 28$) had visual estimates one interval larger than their measured length. A Fisher's exact
499 test comparing the number of vermilion snapper observed in each of the 25 mm size intervals by
500 each survey method was significantly different ($P < 0.001$). However, all visual size estimates
501 assigned the same age as the measured length.

502

503 Discussion

504

505 Annual variation

506

507 There were significant differences detected among years for densities of age-0 vermilion
508 snapper. Densities in August and September 2009 were higher than other years. Similarly, age-
509 0 red snapper also showed high densities on patch-reefs in 2009 (Mudrak and Szedlmayer 2020).
510 This indicates that in 2009 conditions were favorable for recruitment and survival for these two
511 species. Also, densities in October 2014 and 2015 were significantly higher than most other
512 years (except 2010). However, the factors that caused these high densities cannot be determined
513 from the present study.

514 An important question in the present study was if juvenile vermilion snapper densities
515 were affected by the 2010 Deepwater Horizon oil spill. Age-0 densities on the Off-Jul2010 and
516 Off-Aug2010 patch-reefs were similar to densities on the Off-Aug2007 patch-reefs before the oil
517 spill, and not statistically different from the In-Jul2013, In-Jul2014, and In-Jul2015 patch-reefs
518 in some months. All surveys on the In-Aug2010, In-Jul2011, and In-Jul2012 patch-reefs
519 observed a complete absence of juvenile vermilion snapper. However, mean densities of age-0
520 vermilion snapper on the Off-Jul2010 and Off-Aug2010 patch-reefs indicates that 2010 was not
521 a year of complete year class failure. The absence of juvenile vermilion snapper on the In-
522 Jul2010, In-Jul2011, and In-Jul2012 patch-reefs may be an indication of an oil spill effect, or it
523 could be caused by patchy settlement of vermilion snapper in space and time. For example, it is
524 possible that the offshore location offered better habitat or received more settlers than the inshore
525 location. Similarly, in 2011 no age-0 vermilion snapper were observed at the Center site used in
526 interannual comparisons, but low densities were observed on patch-reefs deployed farther to the
527 east. These patterns make it difficult to determine if the low densities observed following the oil
528 spill were caused by the spill, natural variability, or some other unknown variable.

529 Age-0 vermilion snapper appear to show different settlement patterns than the age-0 red
530 snapper and gray triggerfish observed on these same patch-reefs. While red snapper and gray
531 triggerfish densities did vary between years, they were consistently present each year of the
532 study (Mudrak and Szedlmayer 2020; Szedlmayer and Mudrak 2022). Vermilion snapper had
533 greater variability, with large schools present on some patch-reefs in some years, and complete
534 absences in other years. Also, in contrast to red snapper and gray triggerfish, large schools of
535 age-0 vermilion snapper were often observed on nearby larger reef structures. Age-0 red snapper
536 and gray triggerfish typically seek out smaller structures where predation risk and competition

537 are reduced, only moving to larger reef structure when larger sizes are obtained as age-1 or age-2
538 individuals (Gallaway et al. 2009; Mudrak and Szedlmayer 2012).

539 Conversely, there were no significant difference in juvenile vermilion snapper densities
540 between patch reefs placed 15 m from larger reef structure, and the patch-reefs placed 500 m
541 from larger reef structure (Mudrak and Szedlmayer 2012). In fact, there are some indications
542 that vermilion snapper may prefer larger reef structures compared to the present study patch-
543 reefs. For example, Szedlmayer and Brewton (2020) provided a photograph of a larger reef that
544 contained such high densities of age-0 vermilion snapper that the actual reef structure was not
545 visible. If vermilion snapper prefer to settle onto larger reefs such as those inhabited by adults,
546 then patch-reefs may not be the best method to measure juvenile density as most recruits would
547 reside on larger reefs, with patch-reefs only containing the spill over in years of exceptionally
548 high densities. In addition, it is known that adult vermilion snapper prefer deeper water habitats
549 compared to the patch-reef locations in the present study (Jaxion-Harm et al. 2018). If juvenile
550 vermilion snapper have similar preferences future research should examine reefs at deeper depths
551 for newly settled vermilion snapper.

552 Densities of age-1 vermilion snapper were low in the June surveys. This indicates that
553 these fish either suffered high mortality rates, with few individuals surviving over the winter, or
554 that vermilion snapper had emigrated from patch-reefs in search of larger reefs by June. Either
555 way, this represents a difference in life history between juvenile vermilion snapper and juvenile
556 red snapper and gray triggerfish. If vermilion snapper had emigrated rather than suffering
557 mortality, this would be a difference in behavior compared to red snapper and gray triggerfish
558 that had high densities of age-1 individuals on patch-reefs in June (Mudrak and Szedlmayer

559 2020; Szedlmayer and Mudrak 2022). If the decline is in fact caused by mortality, then the
560 mortality rate for juvenile vermilion snapper is higher than other fish residing on the patch-reefs.

561 Divers often had difficulty in counting vermilion snapper, as small age-0 vermilion
562 snapper can be difficult to identify. This is especially true when they are mixed with age-0
563 tomtate and round scad as was often observed. Vermilion snapper also appeared to show diver
564 avoidance, and many vermilion snapper were observed at the edge of visibility away from the
565 patch-reef. Therefore, visual estimates may have underestimated vermilion snapper densities
566 simply due to visibility.

567

568 Age-0 and age-1 vermilion snapper relations

569

570 There were no significant correlations between the density of age-0 and age-1 vermilion
571 snapper in any of the months examined. This is in contrast to red snapper, which showed a
572 negative correlation between age-0 and age-1 densities and gray triggerfish which showed a
573 positive correlation between age-0 and age-1 densities on the patch-reefs examined here (Mudrak
574 and Szedlmayer 2012, 2020a, 2020b; Szedlmayer and Mudrak 2014). It may be that vermilion
575 snapper show no preference for or against habitat inhabited by older conspecifics, but the
576 densities of age-1 vermilion snapper on patch-reefs were low in all months examined and it is
577 possible that age-1 densities were not high enough to affect age-0 densities.

578

579 Other species correlation

580

581 The diverse fish assemblages on these patch-reefs could allow for many correlation tests,
582 especially if those species are then further divided into age classes. To avoid running a large
583 number of tests, some of which would be expected to be significant based on probability alone
584 (type-I error), only correlations with red snapper and age-0 tomtate were analyzed. Red snapper
585 was selected because they were the most abundant species on the patch-reefs and represented a
586 potential competitor or predator of juvenile vermilion snapper. Also, as red snapper stocks
587 recover, it is important to measure their effect on other reef fishes. Juvenile (age-0 and age-1)
588 tomtate were selected for analysis based on observations of juvenile tomtate and vermilion
589 snapper schooling together on the patch-reefs.

590 There was a significant positive correlation between juvenile red snapper and vermilion
591 snapper in September. This likely resulted from exceptionally high densities of both species in
592 September 2009 (Mudrak and Szedlmayer 2020). No other significant relations were detected
593 between the two species. The positive correlation observed in September does not support a
594 competitive exclusion effect by red snapper on vermilion snapper for these early life history
595 stages, as vermilion snapper were able to colonize patch-reefs despite high densities of red
596 snapper.

597 There was also a significant positive correlation between juvenile vermilion snapper and
598 tomtate in October. This positive correlation along with positive correlations between vermilion
599 snapper and other reef fishes (after excluding tomtate and red snapper) suggests that patch reefs
600 with favorable conditions led to higher densities of all reef fish species in October. These patch-
601 reefs may have had better food resources, or perhaps higher densities of resident fishes increases
602 the likelihood that new recruits will locate such a patch-reef.

603

604 Mortality

605

606 In the present study we assumed that after patch-reefs reached maximum densities and
607 declines in abundance were attributed to mortality rather than emigration. The mean density of
608 vermilion snapper declined between the fall and the following summer for most years. However,
609 in 2011 and 2013 densities increased from fall to summer and mortalities could not be estimated.
610 This may possibly be explained by the overall low numbers of juvenile vermilion snapper in
611 2011 and 2013, where the immigration of a small number of individuals over the winter and
612 spring caused higher counts in the summer than were observed the previous fall.

613 The observed mortality rates did not indicate density dependent mechanisms. Similar
614 high mortality rates were observed in years of both higher and lower vermilion snapper densities.
615 It is possible that density independent processes, such as recruitment drive vermilion snapper
616 populations dynamics at the juvenile stage. It is also possible that densities of vermilion snapper
617 observed on the present patch-reefs were too low to evaluate density dependent effects.

618 The observed mortality rate is reported as Z because it represents all sources of mortality.
619 However, this could be assumed to represent M as these fish are below the minimum size limits
620 for the directed fishery, and vermilion snapper are residing on small patch-reefs that are difficult
621 to find and target with hook-and-line making release mortality unlikely. Also, if a trawl passed
622 over the patch-reef the patch-reef would be damaged or lost entirely. Therefore, Z and M
623 estimates were assumed the same.

624 While the open nature of these patch-reefs suggests that mortality estimates should be
625 treated with caution, i.e., emigration may cause mortality to be overestimated, at the time of
626 writing these estimates represent the only mortality estimates of juvenile vermilion snapper.

627 Based on these estimates, a mean M of 3.51 is recommended for the time period shortly after
628 settlement to the following summer. This value is substantially higher than the age-0 and age-1
629 natural mortality rates recommended in SEDAR 67 of $M = 0.234$ and $M = 0.342$ (SEDAR-67
630 2020).

631

632 SEMAP trawl surveys

633

634 In October there were six years available for comparisons, and a significant correlation
635 was detected between the density of age-0 vermilion snapper on patch-reefs and the CPUE of
636 age-0 vermilion snapper in SEMAP trawl tows. Unfortunately, patch-reefs in 2009 that showed
637 the highest age-0 vermilion snapper densities in September were not surveyed in October, and
638 we were unable to make comparisons for that year. Also, an important caveat that needs
639 consideration in the present density estimates is whether age-0 vermilion snapper prefer to settle
640 directly onto larger reef habitats, if so both patch-reef and trawl surveys may be unsuitable for
641 density estimations.

642 There were no significant correlations between age-1 densities on patch-reefs, and trawl
643 CPUE in June. This is likely due to the low densities of age-1 vermilion snapper present on the
644 patch-reefs in June. The density of age-1 vermilion sapper on patch-reefs was often lower than
645 the CPUE in trawl samples. This is in contrast to red snapper and gray triggerfish, which had
646 densities on patch-reefs many times higher than the CPUE in trawl sampling (Mudrak and
647 Szedlmayer 2020; Szedlmayer and Mudrak 2022). There are several possible explanations: 1)
648 patch-reefs were unsuitable habitat for age-1 vermilion snapper, 2) age-1 vermilion snapper
649 avoided the divers and were not counted, or 3) trawl surveys occurred over habitat more suitable

650 for vermilion snapper. This could be deeper water, areas father east, or perhaps the trawls ran
651 close to or over larger reef habitats.

652

653 Spatial distributions

654

655 In 2011, age-0 and age-1 vermilion snapper were only present on the patch-reefs
656 deployed farthest to the east. This may indicate better habitat conditions, or a better supply of
657 settlers for vermilion snapper as one moves from west to east within the present study area.
658 Szedlmayer and Mudrak (2014) indicated that substrates had higher silt content to the west and
659 higher sand content to the east. These coarser sand substrates may be preferred by vermilion
660 snapper. The East site was also a greater distance from the Mississippi River and Mobile River
661 discharges that results in reduced sedimentation and other freshwater influences. In 2011 the
662 West and possibly the Center site were affected by hypoxic conditions in August (Szedlmayer
663 and Mudrak 2014). However, vermilion snapper were completely absent from the West and
664 Center sites before the hypoxic event, indicating that hypoxia was not the cause of their absence.

665

666 2015 Spatial Experiment

667

668 In 2015, it was possible that the 100 patch-reefs deployed at the inshore location since
669 2010 were providing a source of early colonizers to new patch-reefs deployed in the area.
670 Therefore, additional patch-reefs were deployed 11 km to the east to remove this potential source
671 of immigrants. The patch-reefs at the Center site were initially colonized by higher densities of
672 age-1 red snapper and gray triggerfish (Szedlmayer and Szedlmayer 2022). This in turn lead to

673 lower initial densities of age-0 red snapper, which have a negative relation with age-1
674 conspecifics, and higher initial densities of age-0 gray triggerfish, which have a positive relation
675 with age-0 conspecifics (Mudrak and Szedlmayer 2012; Szedlmayer and Mudrak 2014; Mudrak
676 and Szedlmayer 2020; Szedlmayer and Mudrak 2022).

677 This experiment also observed a significant effect on vermilion snapper with higher
678 densities of both age-0 and age-1 individuals on the 2015 East site, which was farther from
679 known sources of immigrants. However, this study did not identify any relation between
680 conspecifics and other reef fish that could easily explain the increased densities of vermilion
681 snapper on the 2015 East site. It is possible that there was an unidentified negative relation
682 between vermilion snapper and these early immigrants of age-1 red snapper or gray triggerfish at
683 the center site. It is also possible that the 11 km distance between the two study sites was far
684 enough to place the 2015 East site in an area with better habitat or a higher supply of settlers for
685 vermilion snapper. For example, in the 2011 spatial experiment, vermilion snapper were only
686 present on the East site. Therefore, while the effect of this 11 km distance to the east was
687 significant, it is unclear what variable caused the effect.

688

689 Removal experiment

690

691 The removal experiment did not detect a significant effect on either age-0 or age-1
692 vermilion snapper densities. Rather than concluding that the density of red snapper and other
693 reef associated fishes had no effect on juvenile vermilion snapper densities, it is more likely that
694 the removal treatments were not successful in effectively lowering fish densities. Also, the

695 removal experiment occurred in a year of low vermilion snapper densities. We suggest that
696 future removal experiments be carried out during years with higher vermilion snapper densities.

697

698 Drop-net-rotenone vs visual estimates

699

700 Drop-net-rotenone samples can underestimate densities because not all individuals on a
701 patch-reef are captured but cannot overestimate densities because it is not possible to capture
702 more individuals than are present. Visual estimates are capable of either overestimating densities
703 by counting individuals twice, or underestimating densities by missing individuals. The Fisher's
704 exact test did find a significant difference between the visual counts and the drop-net catches.
705 However, while drop-nets caught more individuals than were counted in the visual surveys, most
706 of this difference came from a single patch-reef where 24 vermilion snapper were visually
707 counted, and 34 vermilion snapper were caught in the drop-net.

708 Drop-net sampling allowed for the validation of visual size estimates. All individuals
709 counted in visual surveys were within one 25 mm size interval of their measured length. Thus,
710 as long as the size of most age-0 and most age-1 individuals differ by at least 50 mm, few
711 individuals will be assigned to an incorrect age and conclusions based on visual size estimates
712 will be valid.

713

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726

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794 triggerfish (*Balistes capriscus*) based on small artificial patch-reefs. Gulf of Mexico
795 Fishery Management Council Final Report, Tampa, FL.
- 796 Zar, J. H. 2010. Biostatistical Analysis, fifth edition. Prentice Hall, Englewood Cliffs.
- 797

798 Table 1. Environmental conditions associated with visual surveys for juvenile vermilion snapper: Temperature = Temp, salinity = Sal
 799 and dissolved oxygen = DO measured within 1 m of the seafloor during each survey. If more than one measurement was recorded, the
 800 mean value is displayed.

Reef Set	August			September			October			June		
	Temp °C	Sal ‰	DO mg/L									
Off-Aug2007	–	–	–	–	–	–	–	–	–	22.9	34.3	4.5
Off-Jul2008	–	–	–	–	–	–	–	–	–	–	–	–
Off-Jul2009	23.6	29.0	5.7	28.2	29.3	6.8	–	–	–	–	–	–
Off-Jul2010	23.7	32.2	2.4	26.4	33.1	3.8	–	–	–	–	–	–
Off-Aug2010	–	–	–	26.3	33.0	2.4	24.8	33.7	6.5	–	–	–
In-Aug2010	–	–	–	28.2	30.6	2.0	24.0	36.2	5.8	–	–	–
In-Jul2011	25.3	35.5	2.4	–	–	–	24.2	33.3	5.5	–	–	–
In-Jul2012	–	–	–	–	–	–	–	–	–	–	–	–
In-Jul2013	–	–	–	–	–	–	27.9	31.6	–	–	–	–
In-Jul2014	–	–	–	30.0	32.3	–	–	–	–	27.1	33.6	–
In-Jul2015	28.9	34.3	5.6	–	–	–	25.6	32.5	4.5	23.9	34.8	–

801

802 Table 2. Location and deployment date for patch-reef sets surveyed off Alabama, in the northern Gulf of Mexico. Reef sets located
 803 inshore (12 – 16 km) are prefixed with “In”, and reef sets located offshore (19 – 23 km) are prefixed with “Off”. Reef N = the number
 804 of reefs deployed in each reef set. Survey N = number of reefs surveyed for each month (not all reefs deployed were surveyed each
 805 month). Dates of surveys are listed within each month.

Reef Set	Reef N	Reef Deployed	Surveys							
			August	N	September	N	October	N	June	N
Off-Aug2007	30	1-9Aug07	-	-	27Sep07	10	26Oct07	10	10-19Jun08	24
Off-Jul2008	10	24-28Jul08	6-15Aug08	10	-	-	-	-	-	-
Off-Jul2009	10	9-10Jul09	4-6Aug09	10	9-10Sep09	10	-	-	-	-
Off-Jul2010	10	14-15Jul10	2-3Aug10	10	9-20Sep10	10	-	-	-	-
Off-Aug2010	10	25Aug10	-	-	9Sep10	10	21Oct10	10	30Jun11	10
In-Aug2010	10	24Aug10	-	-	8Sep10	10	18Oct10	10	9Jun11	10
In-Jul2011	10	19-20Jul11	29-30Aug11	10	-	-	26Oct11	9	14Jun12	9
In-Jul2012	10	19Jul12	8Aug12	10	25Sep12	6	-	-	-	-
In-Jul2013	10	18Jul-1Aug13	27-29Aug13	10	-	-	30Sep-16Oct13	9	5-17Jun14	10
In-Jul2014	14	22-24Jul14	21-22Aug14	14	8-10Sep14	14	30Sep-2Oct14	14	2-4Jun15	14
In-Jul2015	15	28Jul15	21-28Aug15	15	-	-	30Sep-7Oct15	15	13-22Jun16	14

806

807 Table 3. Juvenile vermilion snapper total instantaneous mortality (Z) observed each year.
 808 Mortalities were based on the decline between the maximum density of age-0 vermilion snapper
 809 observed on patch-reefs in the fall surveys and the density of age-1 vermilion snapper observed
 810 on the first summer survey the following year. Mortality was not calculated for 2011 and 2013
 811 because densities increased between fall and the following summer. The number of patch-reefs
 812 with available data each year = N . Mean Z was based on the mean for all years.

Year	N	Z
2007	28	3.40
2010	25	4.76
2011	21	
2013	10	
2014	14	2.69
2015	29	4.61
Mean	–	3.87

813

814

815

816 Table 4. Mean CPUE \pm SE (catch/H) of age-0 and age-1 vermilion snapper and the total number
 817 of trawl tows conducted by SEAMAP trawl surveys by year east of the Mississippi River. Only
 818 years with corresponding visual estimates of juvenile vermilion snapper on patch-reefs were
 819 compared.

Year	Season	Age	Mean CPUE	Trawl <i>N</i>
2007	Fall	0	0 \pm 0	34
2010	Fall	0	0.54 \pm 0.36	93
2011	Fall	0	0 \pm 0	17
2013	Fall	0	0.70 \pm 0.49	76
2014	Fall	0	1.41 \pm 0.44	193
2015	Fall	0	1.81 \pm 0.51	142
2008	Summer	1	0.18 \pm 0.13	49
2011	Summer	1	3.49 \pm 0.69	156
2012	Summer	1	3.35 \pm 0.76	140
2014	Summer	1	1.76 \pm 0.49	204
2015	Summer	1	1.99 \pm 0.43	215
2016	Summer	1	3.12 \pm 0.59	141

820

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823 List of Figures

824

825 Figure 1. Small patch-reef deployed in the present study, off coastal Alabama, U.S., in the
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827

828 Figure 2. Locations of artificial patch-reefs deployed and visually surveyed on the Alabama
829 continental shelf in the northern Gulf of Mexico from 2007 to 2015. All symbols represent sets
830 of patch-reefs containing 10 to 30 individual patch-reefs placed at least 500 m apart. Reef set
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832 = 30) deployed at the offshore location in August 2007.

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835 Densities observed prior to 1 January represent age-0 vermilion snapper, while densities
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837

838 Figure 4. Mean densities (number/m³) of (A) age-0 and (B) age-1 vermilion snapper on patch-
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840 age-0 and age-1 densities were analyzed separately. Error bars = SE.

841

842 Figure 5. Mean densities (number/m³) of (A) age-0 and (B) age-1 vermilion snapper on patch-
843 reefs in September for each year. Different letters indicate significant differences ($P < 0.05$), and
844 age-0 and age-1 densities were analyzed separately. Error bars = SE.

845

846 Figure 6. Mean densities (number/m³) of (A) age-0 and (B) age-1 vermilion snapper on patch-
847 reefs in October for each year. Different letters indicate significant differences ($P \leq 0.05$), and
848 age-0 and age-1 densities were analyzed separately. Error bars = SE.

849

850 Figure 7. Mean densities (number/m³) of age-1 vermilion snapper on patch-reefs in June for
851 each year after the patch-reefs were deployed. Different letters indicate significant differences
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853

854 Figure 8. Comparison of mean density (number/m³) of age-0 vermilion snapper on patch-reefs
855 in October to mean CPUE (catch/H) of age-0 vermilion snapper from SEAMAP trawl surveys in
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857

858 Figure 9. Comparison of mean density (number/m³) of age-1 vermilion snapper on patch-reefs
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865 was not included in analysis due to small sample size.

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868 West, Center, and East sites for patch-reefs built in July 2011. No significant differences were
869 observed ($P \geq 0.05$).

870

871 Figure 12. Mean density \pm SE of the 2015 year class (age-0 in 2015, age-1 in 2016) of vermilion
872 snapper on the Center and 2015 East sites for patch-reefs built in July 2015. Different letters
873 indicate significant differences ($P \leq 0.05$).

874

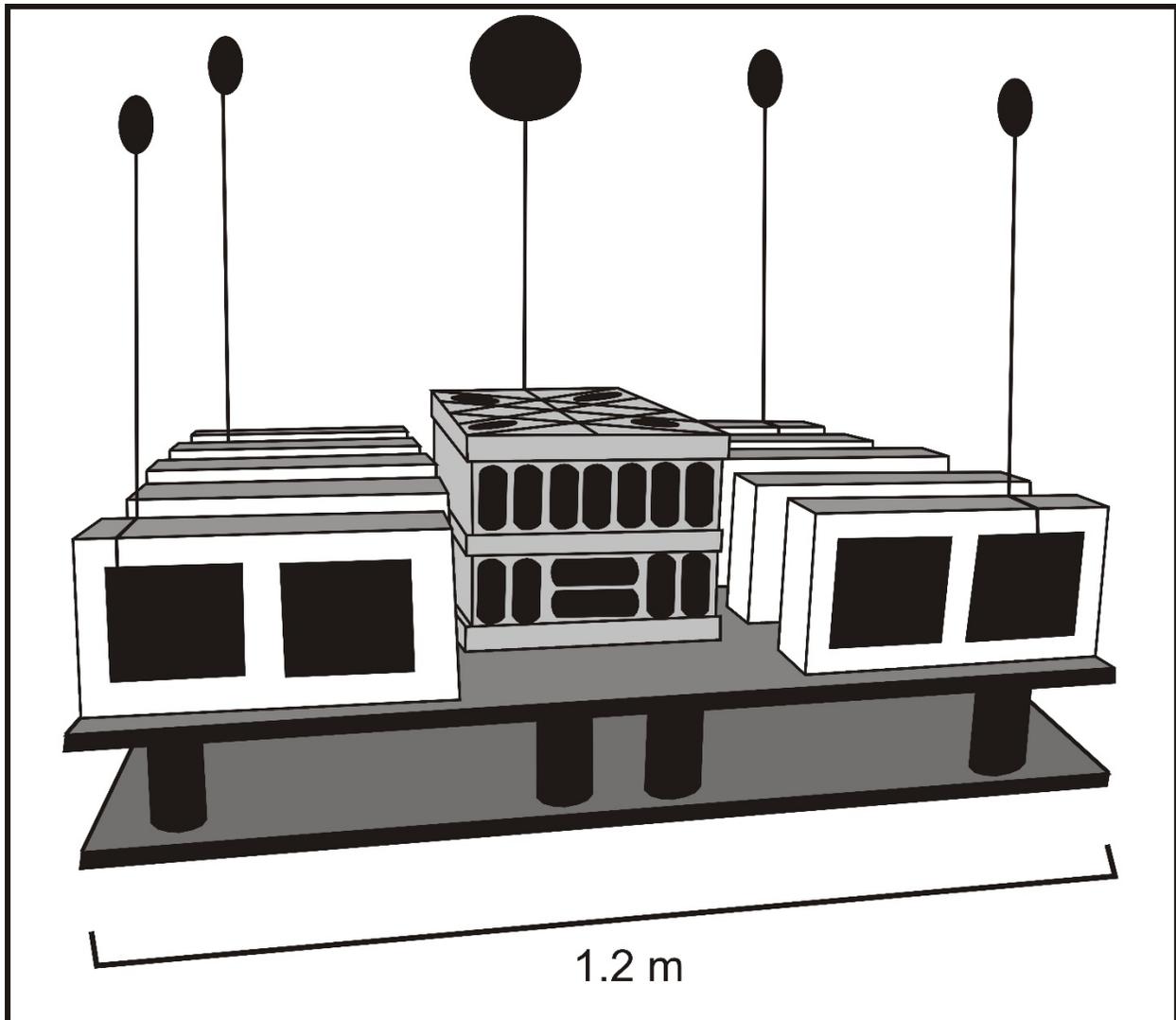
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881 resident fish removed in July (All removal), with only red snapper removed in July (RS
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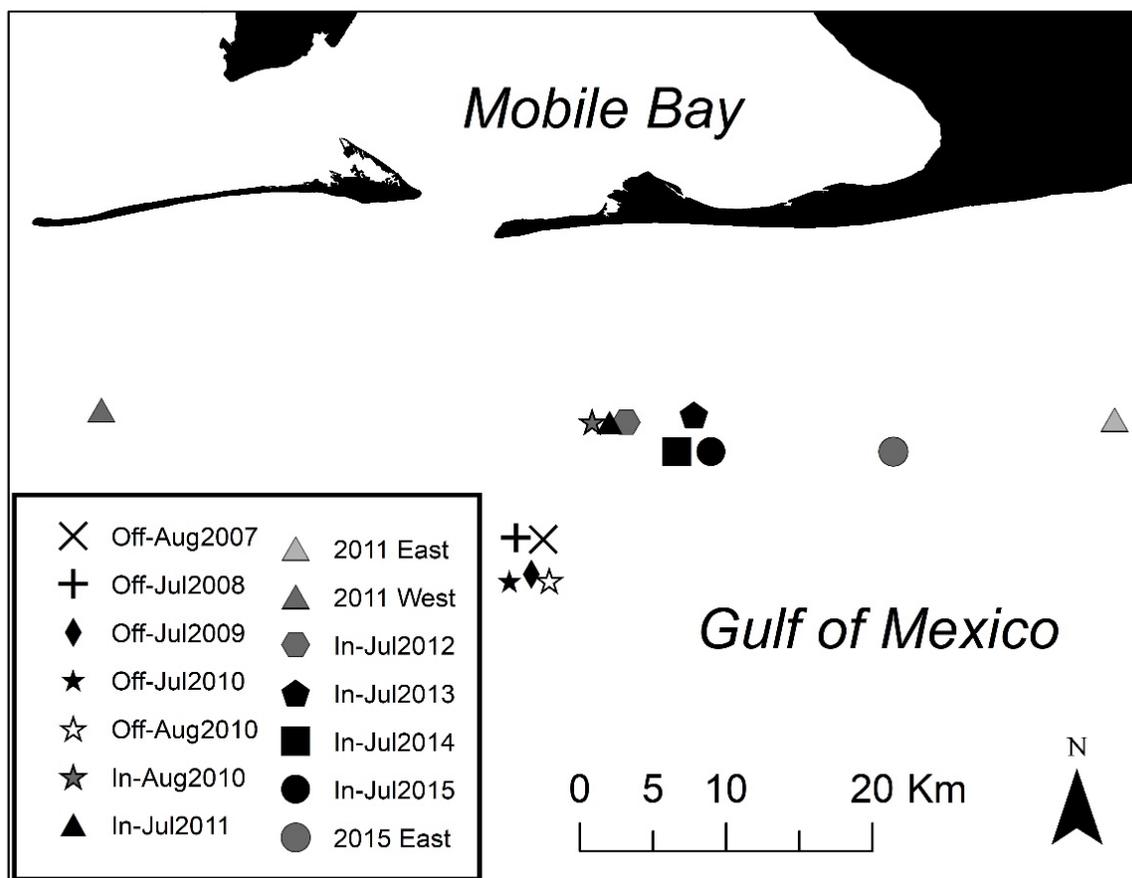
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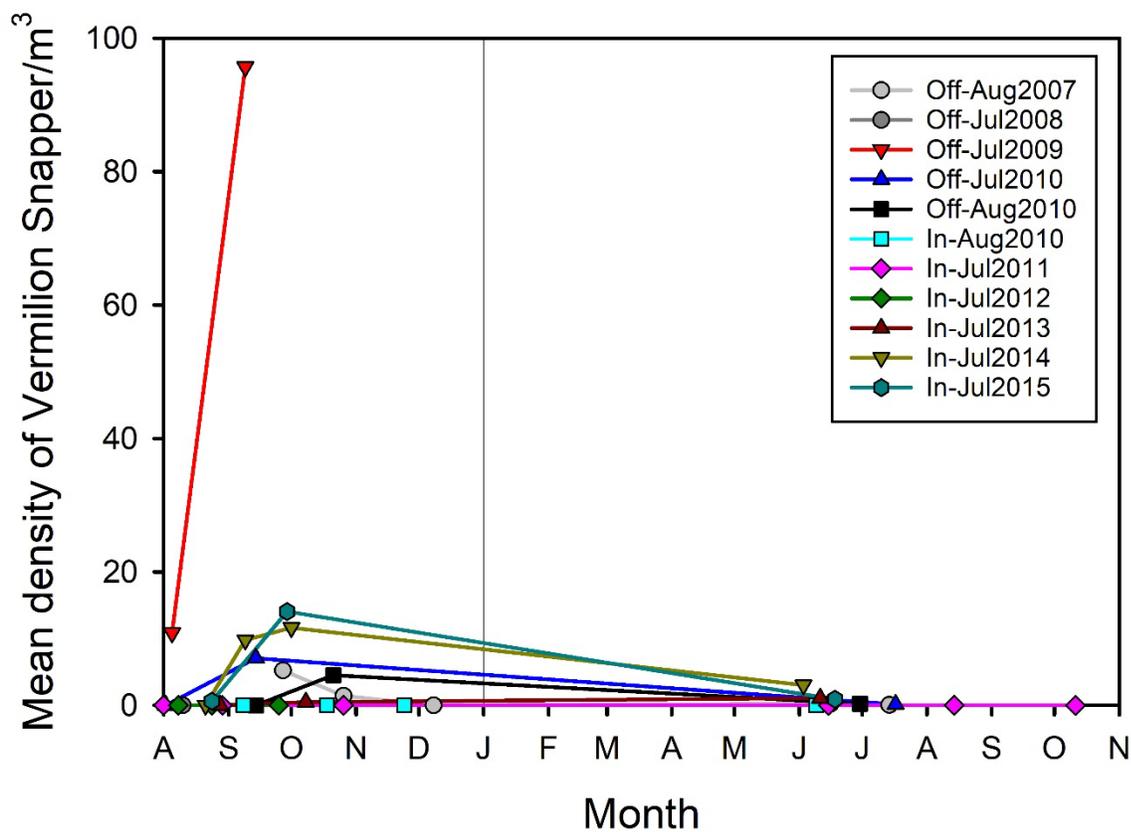
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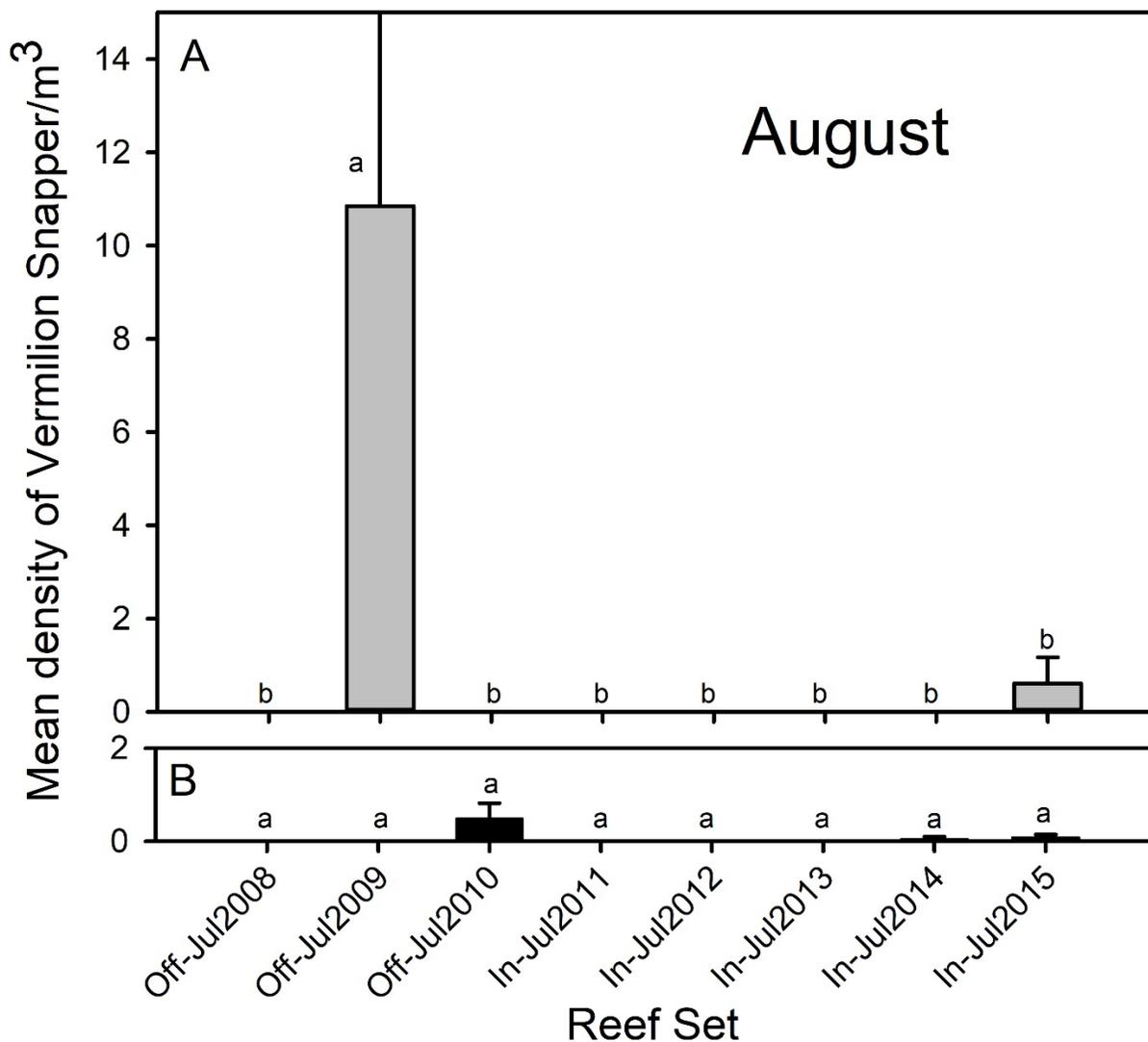
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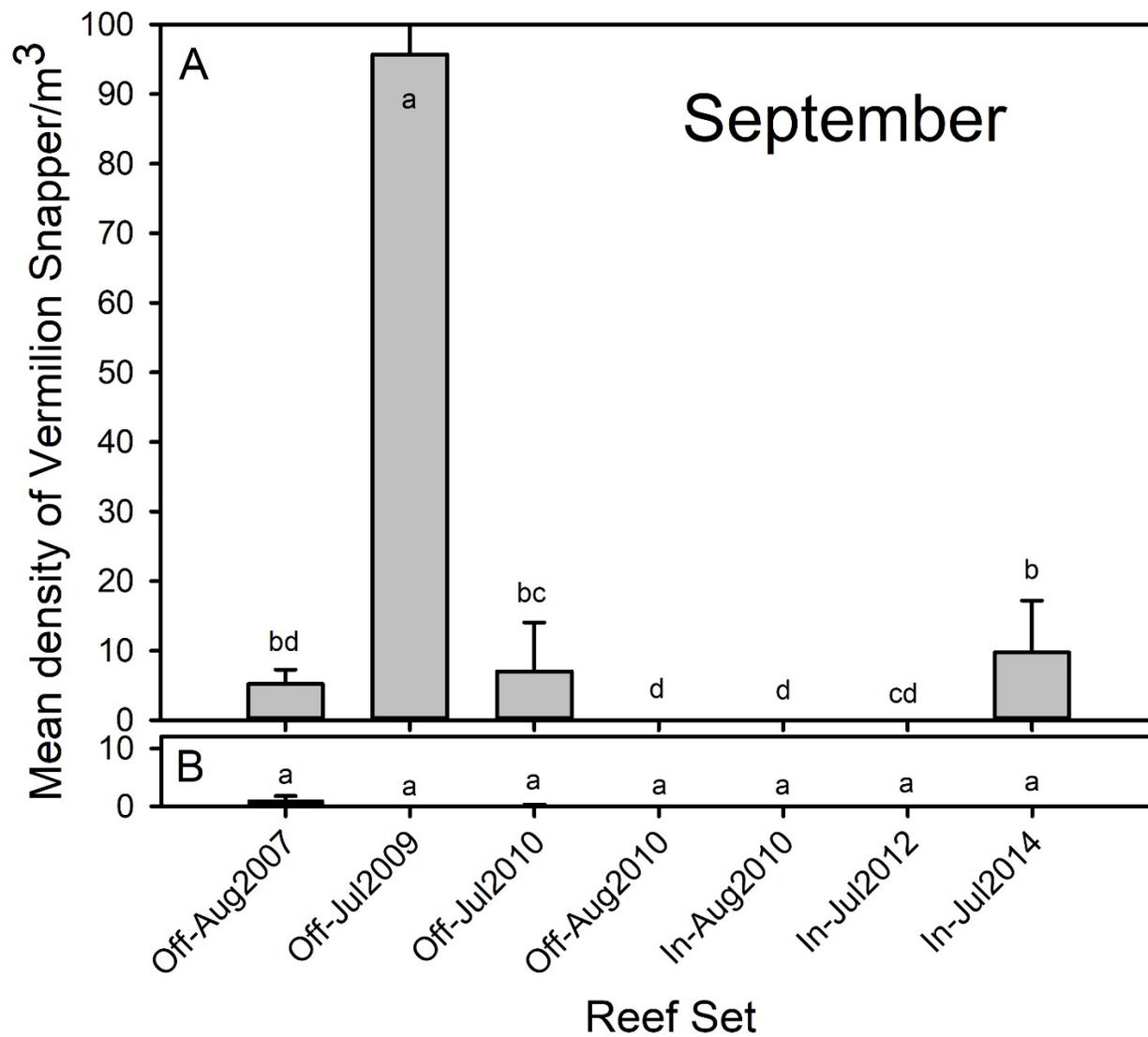
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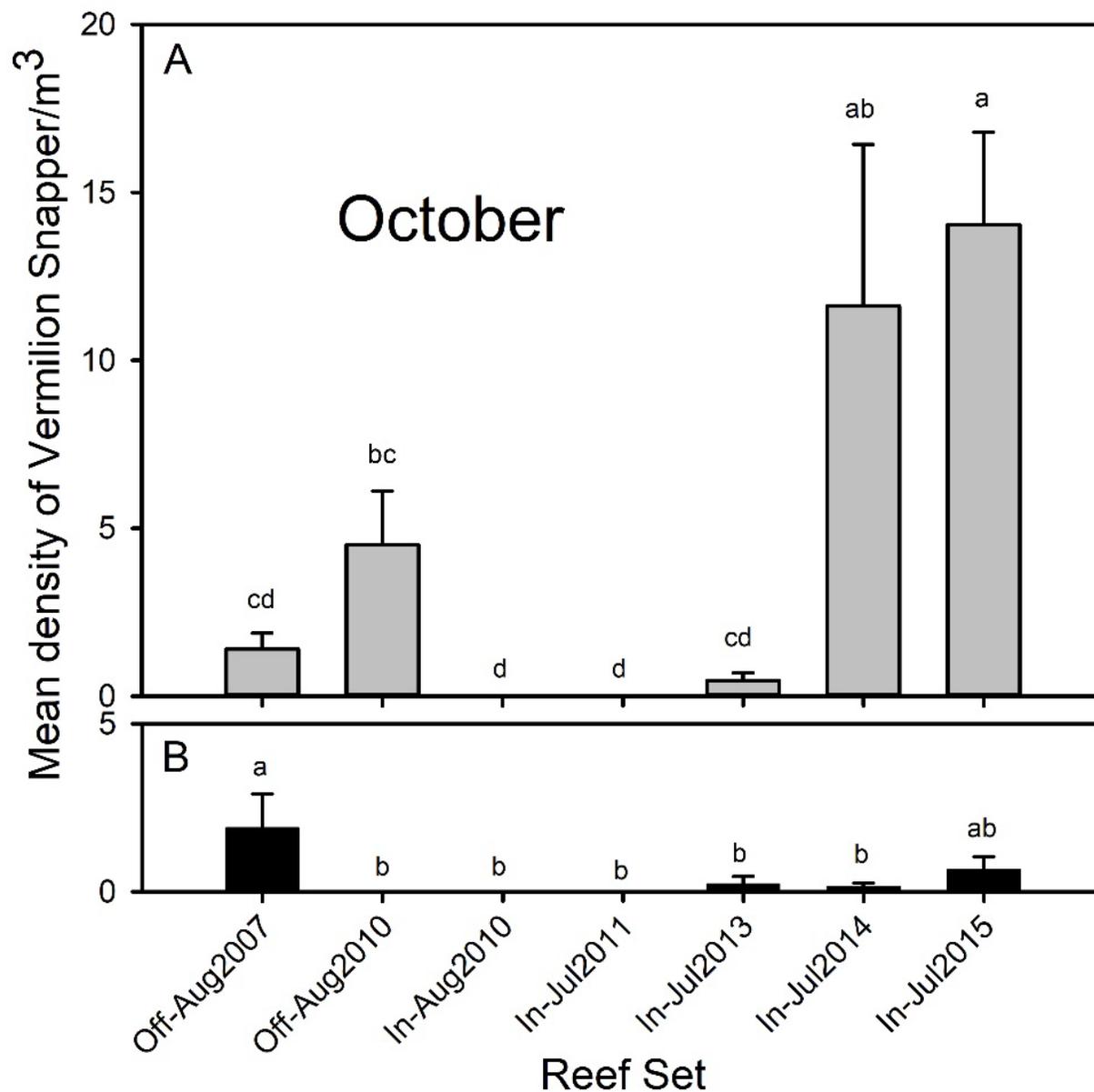
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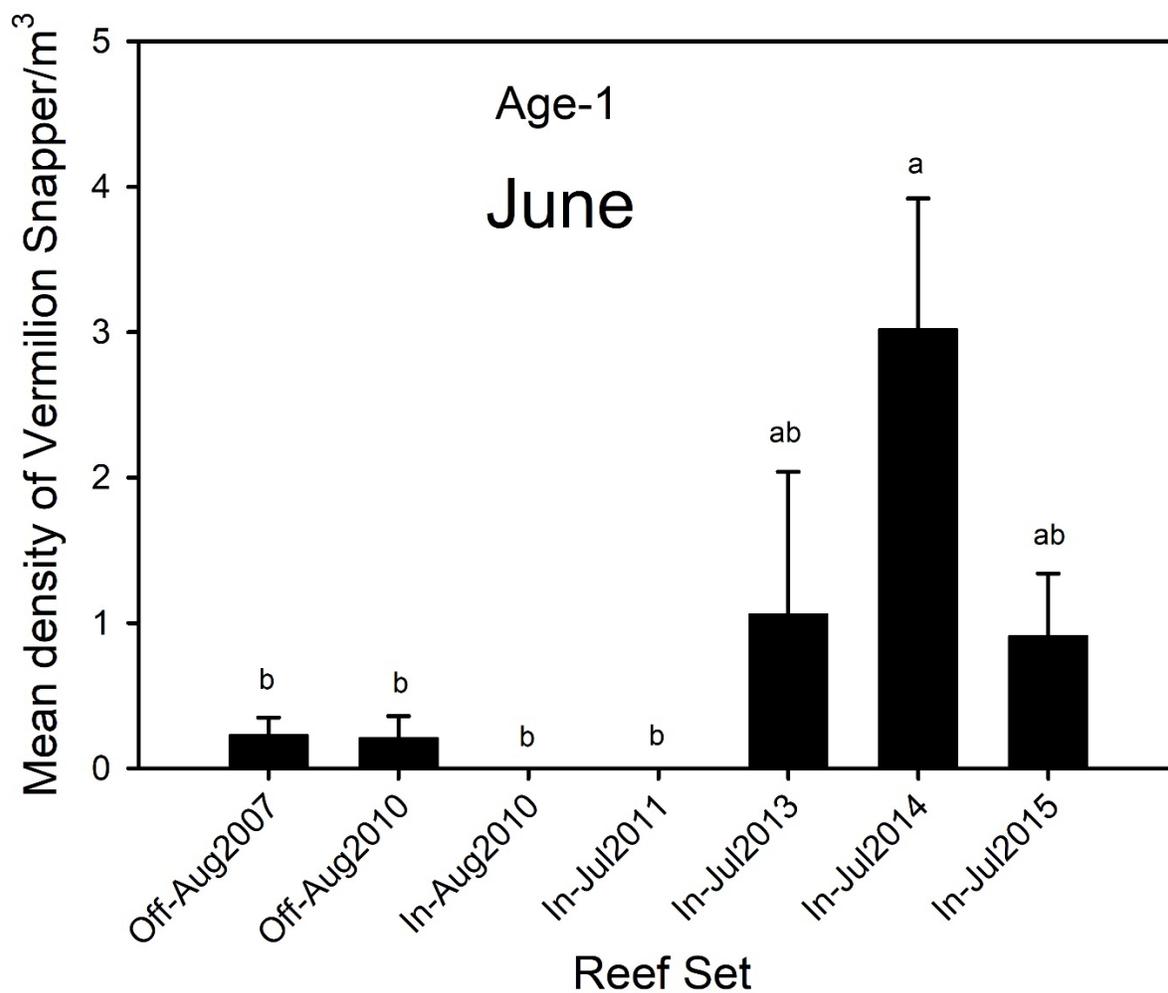
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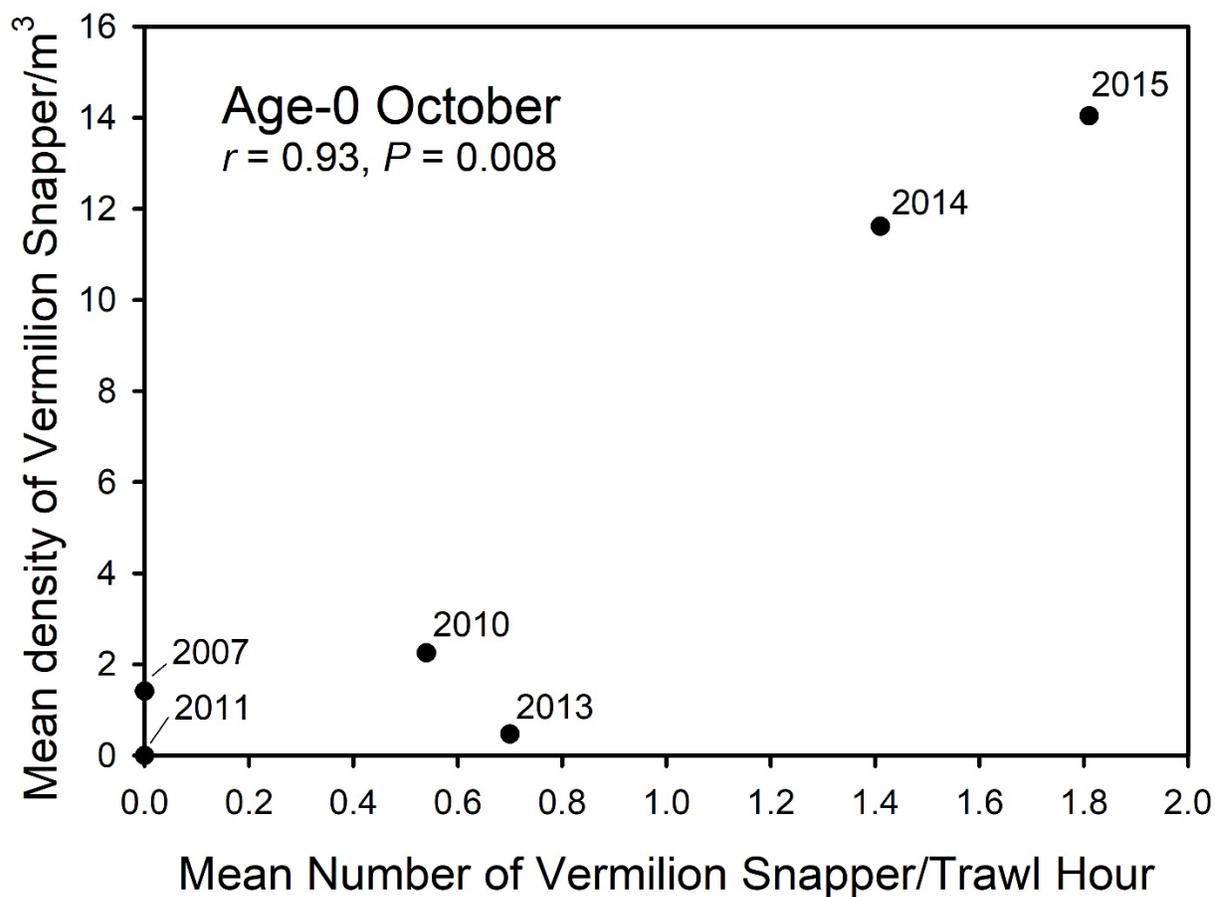
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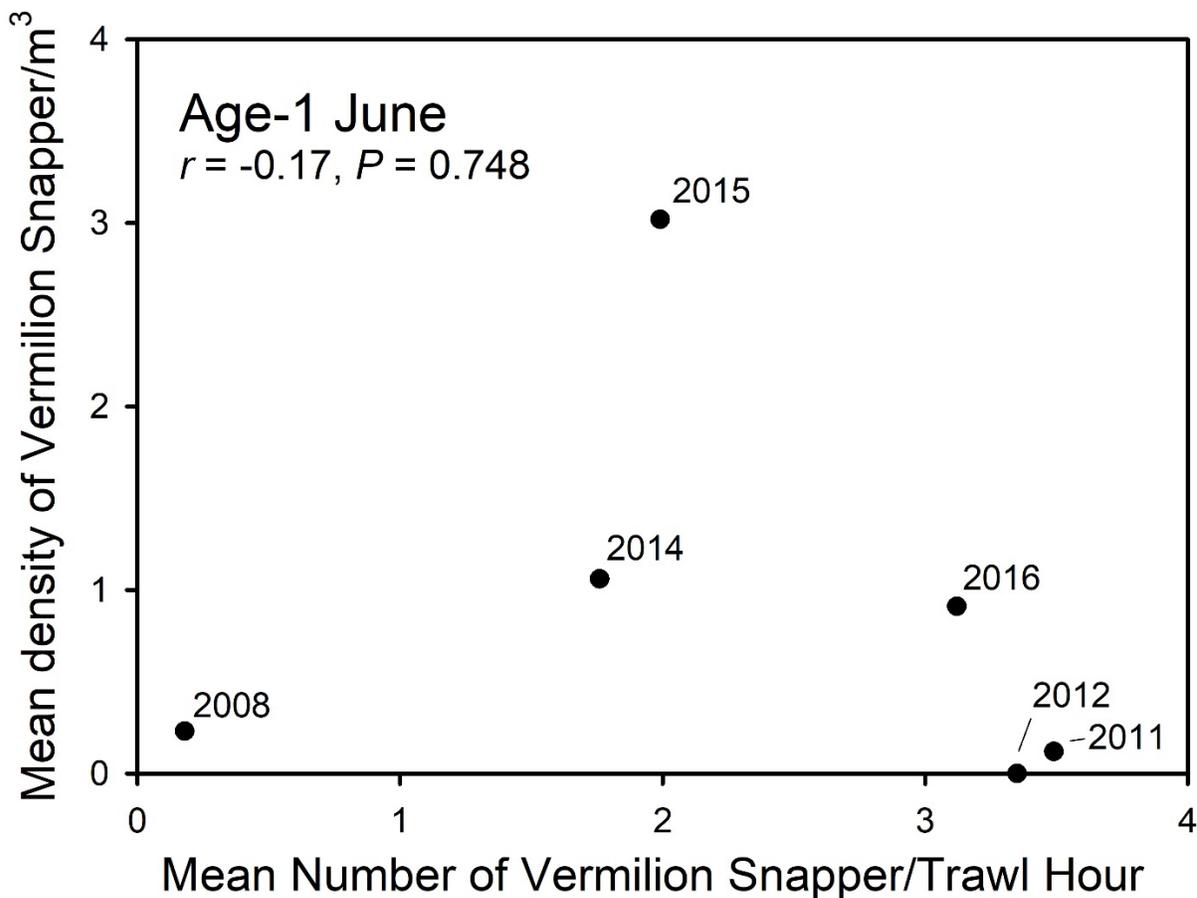
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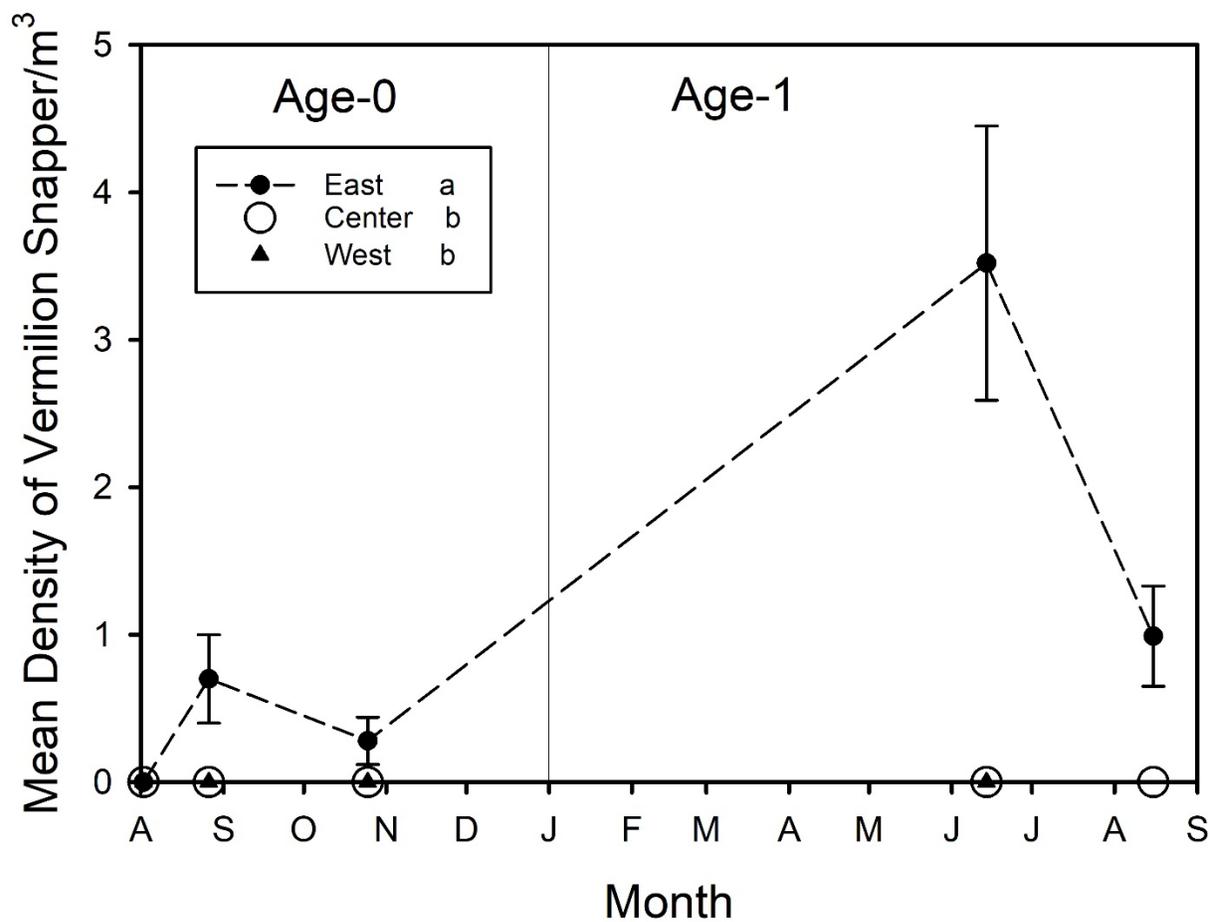
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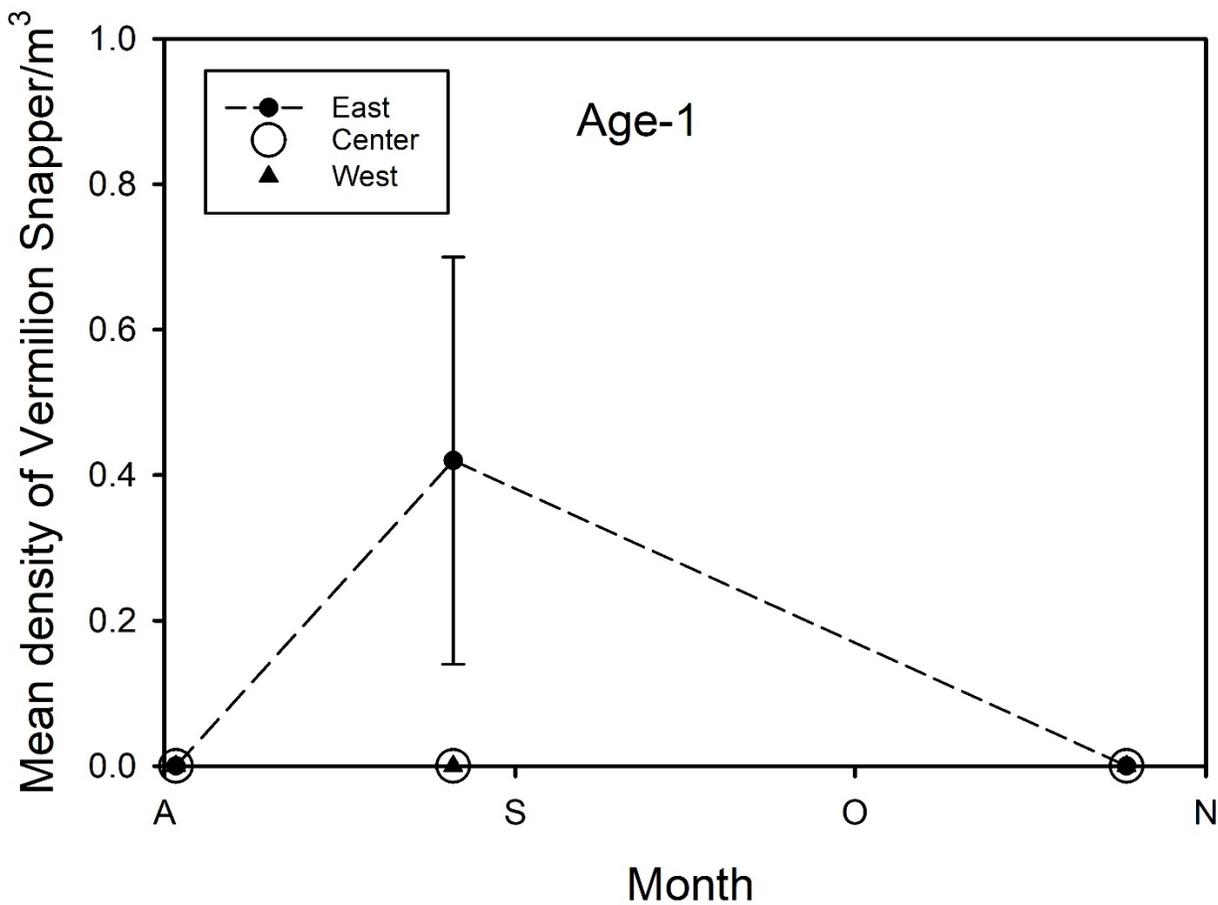
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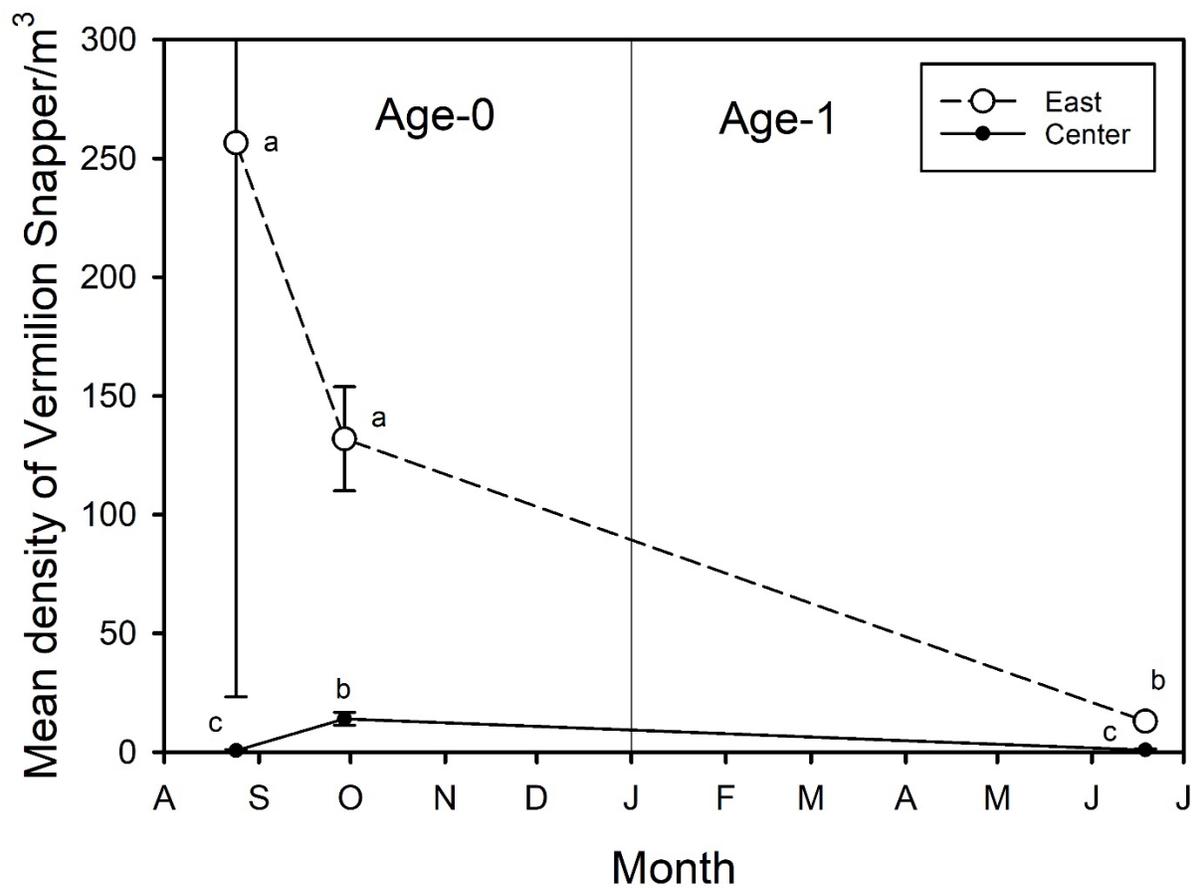
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947



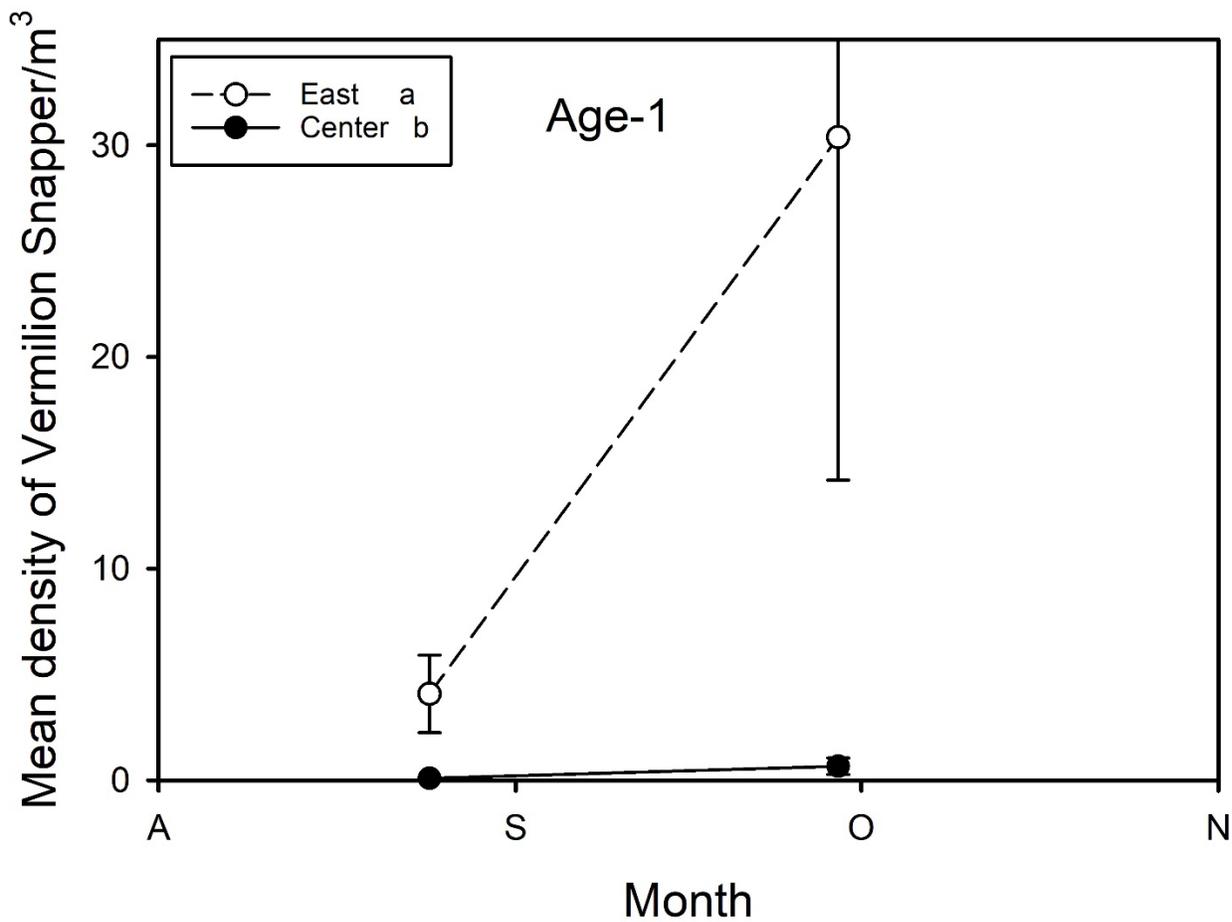
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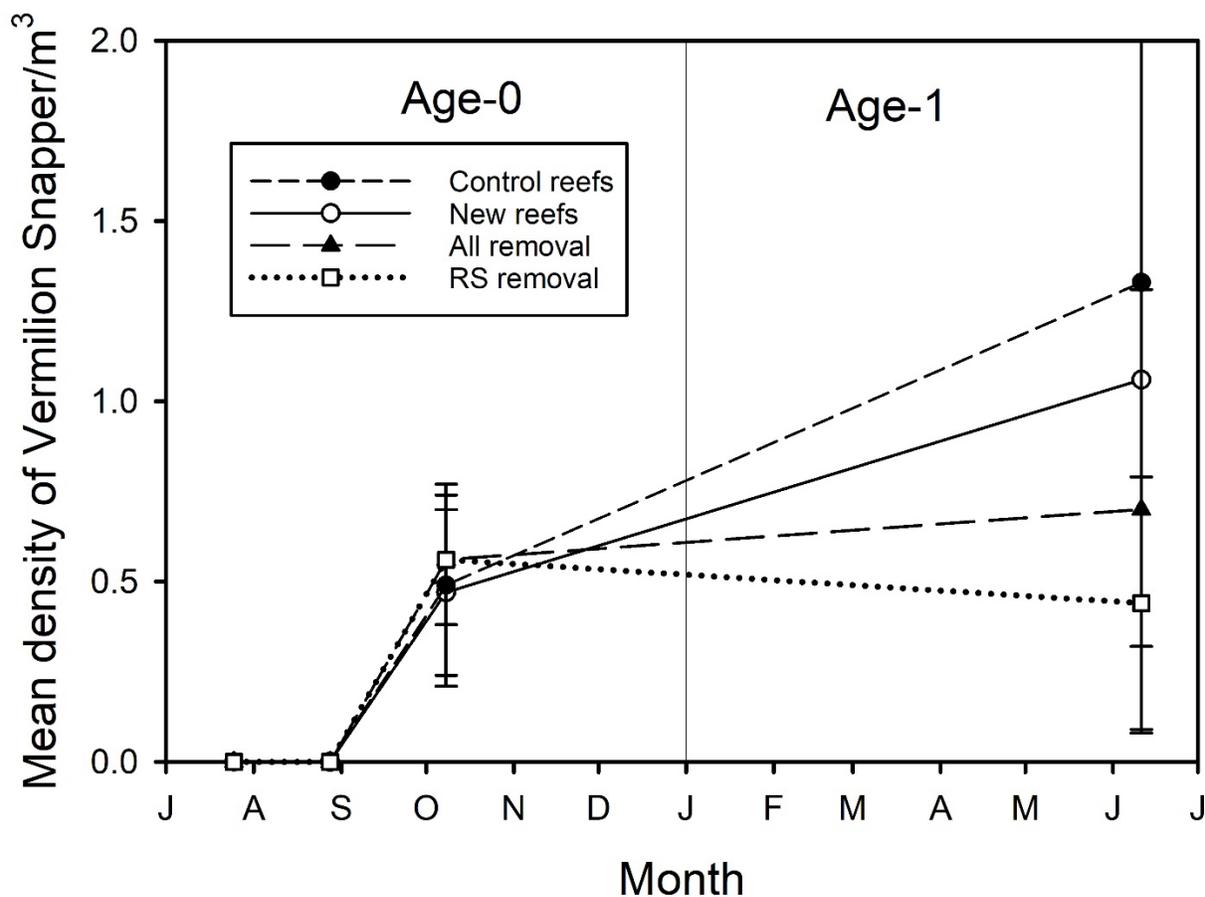
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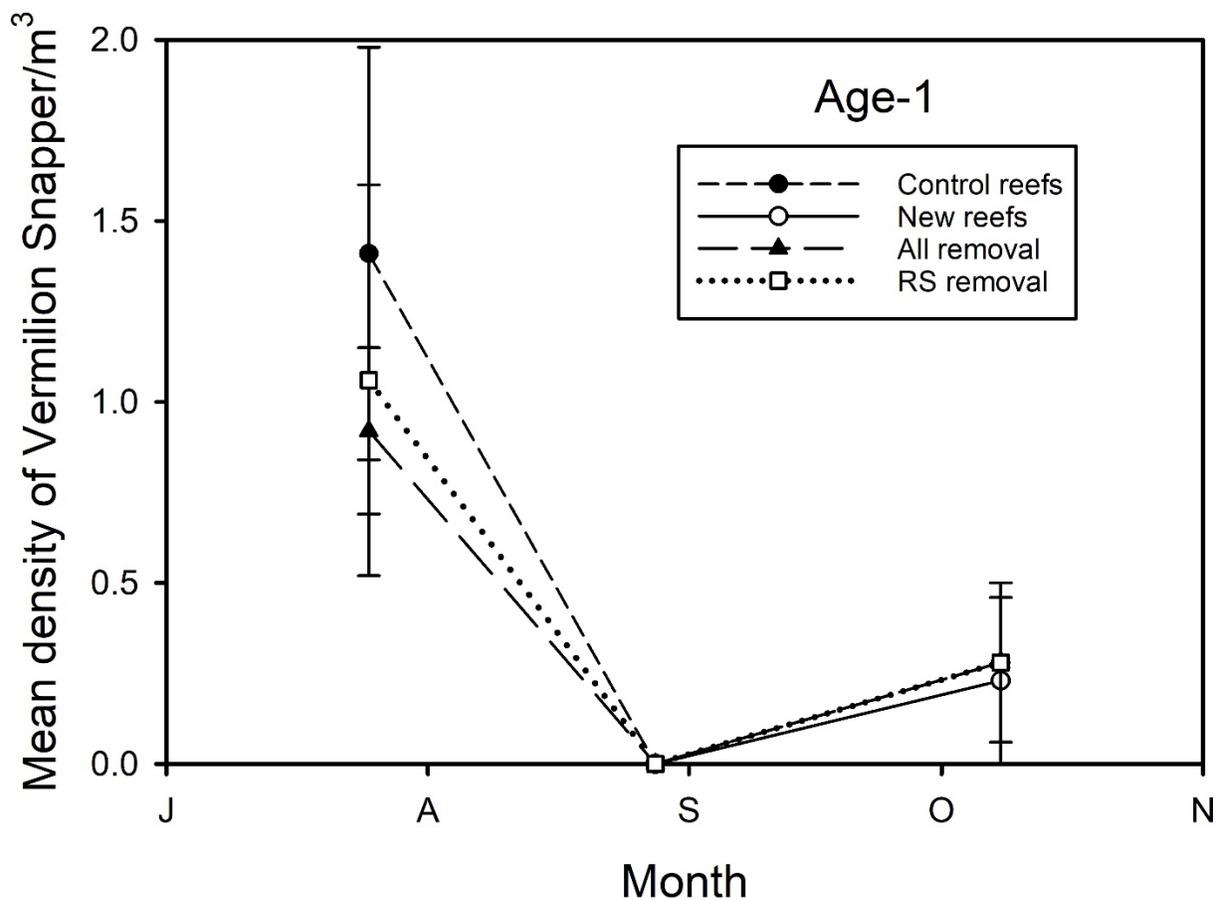
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